

MEPS_2021_Charlotte_Montime
r_Jamaluddin_Jompa_Abdul_Ha
ris.pdf
by

Submission date: 02-Sep-2021 01:12PM (UTC+0700)

Submission ID: 1639942162

File name: MEPS_2021_Charlotte_Montimer_Jamaluddin_Jompa_Abdul_Haris.pdf (1.57M)

Word count: 13570

Character count: 69127



Estimates of sponge consumption rates on an Indo-Pacific reef

Charlotte Mortimer¹, Matthew Dunn², Abdul Haris³, Jamaluddin Jompa³, James Bell^{1,*}

¹School of Biological Sciences, Victoria University of Wellington, Wellington 6012, New Zealand

²The National Institute of Water and Atmospheric Research (NIWA), Wellington 6021, New Zealand

³Universitas Hasanuddin, Department of Marine Science, Makassar 90245, Indonesia

ABSTRACT: Determining predator diets is essential for understanding the strength of top-down processes and how they cascade through food webs. This is especially important for sponges, key members of benthic communities, whose dominance has increased in recent years on some coral reefs. However, the diversity of spongivorous fishes and the sponges they consume are relatively unknown. Here, we estimated sponge consumption by spongivorous fishes in the Wakatobi Marine National Park, Indonesia. We deployed cameras to identify fish biting at the dominant reef sponge *Xestospongia* spp. and then used gut content analysis and fish abundance estimates to quantify sponge consumption. In total, 33 species from 10 families of reef fish were identified taking bites from *Xestospongia* spp.; however, the 2 most prolific sponge-grazers, *Ctenochaetus binotatus* and *Chaetodon kleinii*, had no sponge in their guts, showing that for some fish, bites on sponge surfaces are not reliable evidence of sponge consumption. Gut contents indicated that *Pygospio diacanthus* was an obligate spongivore, while *Pomacanthus imperator*, *P. xanthurus*, *Zanclus cornutus* and *Siganus punctatus* regularly consumed sponges. Sponge consumption by these 5 spongivores was estimated at 46.6 ± 18.3 g sponge $1000\text{ m}^{-2}\text{ d}^{-1}$. Molecular approaches developed to sequence the 18S gene for sponges consumed by angelfishes led to the successful amplification of 14 consumed sponges representing 6 orders of Porifera. We provide the first estimate of sponge consumption in the Indo-Pacific and are the first to successfully sequence partially digested sponges from fish stomachs, identifying several sponges previously unknown to be consumed by spongivores.

KEY WORDS: Spongivory · Angelfishes · Diet · DNA gut contents

Resale or republication not permitted without written consent of the publisher

1. INTRODUCTION

Accurate quantification of a species' functional role is essential to estimate its contribution to ecosystem function (Bellwood et al. 2019). On tropical coral reefs, with the exception of well-studied fish groups such as herbivores and corallivores (Bellwood & Choat 1990, Bellwood et al. 2004, Pratchett 2005, Cole et al. 2011), our understanding of the range of functional roles performed by individual species remains limited, impeding our ability to make informed decisions on ecosystem management (Bellwood et al. 2019). Heightened interest in the interac-

tions between sponges and other organisms (Bell et al. 2018a, Pawlik et al. 2018) has been driven by evidence that sponges may, at least initially, be less affected by this period of unprecedented environmental change than some other benthic organisms (Bell et al. 2013, 2018a,b, Pawlik et al. 2018). Experimental studies indicate that some sponges may have a greater tolerance to environmental shifts than calcifying fauna (Vicente et al. 2016, Bennett et al. 2017). Furthermore, several studies have reported recent increases in sponge abundance (Ruzicka et al. 2013, McMurray et al. 2015, de Bakker et al. 2017), with transitions to sponge-dominated states being report-

*Corresponding author: james.bell@vuw.ac.nz

ed in the Caribbean, Atlantic, Indo-Pacific and Pacific (reviewed in Bell et al. 2013). Increasing sponge abundance will likely have profound impacts on food chain dynamics (Bell 2008, Bell et al. 2013, 2018b), but our ability to predict these changes is restricted by limited data on the taxonomic breadth of spongivores and their preferred prey (Wulff 2017). These knowledge gaps need to be addressed to better understand the future functioning of sponge-dominated reefs (Bell et al. 2018a).

Sponges have structural and chemical defences that deter predation (Hay 1996, Hill et al. 2005, McClintock et al. 2005), but they are thought to rely mainly on secondary chemistry (Pawlik et al. 1995, Pawlik 2011). On Caribbean reefs, angelfishes (Family Pomacanthidae) are well-known spongivores that are thought to reduce the impact of chemical substances produced by sponges by adopting a 'smorgasbord' approach to foraging (Randall & Hartman 1968), whereby small amounts of many sponge species are consumed (Wulff 2006). For example, Randall & Hartman (1968) recorded >40 species of sponge in the gut contents of *Holocanthus ciliaris* and >20 species from other spongivorous angelfishes. Applying existing data on sponge palatability (determined by feeding choice experiments using pelleted extractions of secondary metabolites) to gut content data from Randall & Hartman (1968) demonstrated that the majority of sponges consumed by the angelfishes *Pomacanthus paru* and *P. paru* were chemically undefended (Pawlik et al. 1995, 2018, Pawlik 2011). Two palatable species (*Callyspongia vaginalis* and *Niphates erecta*) made up approximately one-third of gut contents of the 2 angelfish, suggesting a preference for these 2 species in particular (Pawlik et al. 2018).

Indo-Pacific reefs exhibit some of the highest sponge diversity in the world (van Soest et al. 2012); however, few studies have considered the ecological role of spongivory (but see Powell et al. 2014, 2015). In a study conducted in the Wakatobi Marine National Park (MNP), Powell et al. (2014) found that spongivore abundance was highest at Sampela reef, a site that has undergone a coral-sponge regime shift (Powell et al. 2014, Biggerstaff et al. 2015). However, in the study by Powell et al. (2014), sponges were identified from the literature and from *in situ* observations of feeding behaviour, with only 11 of the 21 species listed having gut content data verifying sponges were an ingested prey item. Importantly, fish nibble at sponges but have no predatory impact (Bell et al. 2017, Pawlik et al. 2018), which seems likely for 2 surgeonfishes included in the list of

21 spongivores compiled by Powell et al. (2014) as they likely lack the necessary dental morphology to remove sponge tissue (Bellwood et al. 2014). Combining species that likely fulfill different functional roles on reef systems into a single spongivore category could inadvertently bias our understanding of reef functioning (Wulff 2017). Hence, when the diet of a species remains ambiguous, a multi-method approach can prevent functional misclassifications and provide more detail on a specific diet (Nagelkerken et al. 2009, Bell et al. 2017, Nielsen et al. 2018).

Whilst gut content data provide an unambiguous record of consumed prey (Wulff 2006), it may be biased towards items with structural components that are not easily digested, and taxonomic resolution can be hindered by digestion (Nielsen et al. 2018). Sponges range from having highly dense to very diffuse tissue, which likely affects digestibility and hence tissue available for identification. Over the past decade, molecular methods have emerged as a valuable tool to overcome some of the challenges associated with diet identification (see Nielsen et al. 2018). Polymerase chain reaction (PCR) can detect small amounts of DNA, providing greater resolution to diets that are hard to visually characterise because they lack hard parts, are heavily degraded or are liquid remains (Symondson 2002, Vestheim & Jarman 2008, Nielsen et al. 2018). However, gut content DNA can be degraded and fragmented by digestion (Deagle et al. 2009), and PCR can be inhibited by substances found in digestive tracts (Schradler et al. 2012). Additionally, gut contents contain a mixed profile of samples, and predator DNA may be preferentially amplified over degraded prey DNA (Vestheim & Jarman 2008). Where group-specific primers are not an option, methods that block predator DNA from amplification, e.g. predator-blocking primers (Vestheim & Jarman 2008, Leray et al. 2013, Leray et al. 2015) can be developed to successfully resolve the diets of marine animals.

Here, we aimed to improve our understanding of the ecological significance of spongivory in the Central Indo-Pacific. Our specific objectives were (1) to identify sponge-grazing fishes using camera record fish taking bites from *Xestospongia* spp., one of the most abundant and conspicuous reef sponges in the Wakatobi MNP; (2) to determine diet composition of key sponge-grazing fish using gut content analysis; and (3) to estimate sponge consumption rates using the mean weight of sponge in the gut contents of each spongivore and spongivore density data. Ingested sponges recovered from gut contents could not be easily identified due to a lack of reliable

field guides and a species-rich environment (Bell & Smith 2004, Rovellini et al. 2019). Additionally, homogenisation of sponges within the digestive tract limited traditional identification methods using spicules; therefore, our final objective was (4) to develop molecular methods to identify ingested sponges from the gut contents of spongivorous angelfish from the Indo-Pacific.

2. MATERIALS AND METHODS

2.1. Study sites

This study was conducted at 4 sites within the Wakatobi MNP, the third-largest MNP in Indonesia (Clifton & Unsworth 2013) (Fig. 1). Four sites with different benthic cover, topographical and environmental conditions and fish assemblage compositions (Curtis-Quick 2013, Powell et al. 2014) were chosen to maximise the number of different reef fish species that could be recorded as potential spongivores. All sites were separated by at least 3 km, aside from

Sampela and B1, which are separated by a channel that is approximately 1.5 km wide and extends to a depth of 40 m. Sampela reef was the most degraded of the 4 sites, with sedimentation rates reported as 4 times higher than the other 3 sites (Crabbe & Smith 2005) and with the lowest abundance of live coral cover (~10%) and high abiotic cover (<50%) (Mortimer 2020). This site has experienced a coral-sponge regime shift and has a high abundance of the photophilic encrusting sponge *Lamellodysidea herbacea* (Powell et al. 2014, Biggerstaff et al. 2015, Mortimer 2020). KDS, B1 and The Ridge both had higher coral cover (~30%; Mortimer 2020, Operation Wallacea unpubl. data) but The Ridge had a higher percentage of soft coral (~13%) than Sampela and B1 (~3%; Mortimer 2020). Sponge percentage cover was similar between Sampela, B1 and The Ridge, covering on average $11.9 \pm 1.4\%$ of the substrate. The Ridge had the highest number of standardised sponge species recorded ($9.2 \pm 1.3 \text{ m}^{-2}$), followed by B1 ($6.2 \pm 0.8 \text{ m}^{-2}$) and Sampela ($5.3 \pm 0.5 \text{ m}^{-2}$); no data were recorded for KDS (Mortimer 2020).

2.2. Videos of sponge-grazing fishes

To identify sponge-grazing fish that would be targeted for gut content analysis, videos of the giant barrel sponge *Xestospongia* spp. were conducted at all 4 sites from June to August 2016. We use the term 'sponge-grazing' to describe feeding behaviour whereby fish bite at sponge surfaces but biomass removal (i.e. spongivory) has not been verified. Barrel sponges in the Wakatobi MNP are considered to comprise a species complex including *X. testudinaria*, *X. bergquistia* and at least one undescribed species (Bell et al. 2014). *Xestospongia* spp. was chosen because an initial trial suggested this genus incurs significant grazing activity and was present in sufficient densities across the 4 study sites (McGrath 2018). Additionally, the Caribbean congener *X. muta* was used by Dunlap & Pawlik (1998) to monitor predation by parrotfishes, suggesting that members of this genus attract predators in a variety of reef locations. A GoPro camera was deployed approximately 1 m away from the sponge and set to record from the end of diving activities to maximise the amount of footage in which diving activities did not disrupt fish behaviour (Emslie et al. 2018). To avoid any effect of depth or microhabitat on fish assemblages (Depczynski & Bellwood 2004, Stefanoudis et al. 2019), sponges were filmed on the open reef slope (8–14 m, depending on the site). At the beginning of the video, a measuring

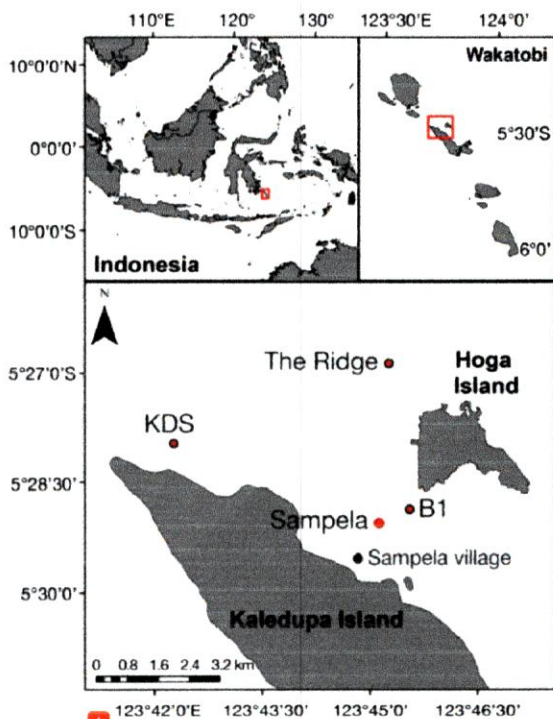


Fig. 1. Map of the 4 study sites and the position of the Hoga/Kaledupa channel within the Wakatobi Marine National Park (MNP) and Indonesia

stick was placed next to the sponge to generate screenshots from which the 2-dimensional area of sponge captured in the frame was estimated to standardise grazing intensity by area. All cameras were deployed during the morning (~08:00–11:00 h) and collected on a subsequent dive in the afternoon, often yielding >2 h of uninterrupted footage.

2.3. Video analysis

The first 5 min of video footage were discarded to account for disturbance from divers leaving the site. The remaining footage was analysed by counting all bites made by fish on *Xestospongia* spp., with bites determined as individual 'strikes' on the sponge surface. Fish were identified to species level except for *Ctenochaetus* spp., as it was not always possible to distinguish between *C. binotatus* and *C. striatus*. Distinguishing individual fish that may have fed repeatedly from the same sponge was not possible using these methods; however, the aim was to measure the grazing intensity of each species rather than individual bite rates. To calculate the area of sponge captured in video footage, screenshots generated from footage containing the scaling device were imported into ImageJ (Abramoff et al. 2004).

2.4. Sponge-grazing fish densities

Sponge-grazing fish densities were quantified at all 4 sites by underwater visual census (UVC) using 6 replicate 50 × 5 m belt transects deployed on the reef slope at 10 m depth, with each transect separated by 15 m. For each transect, the lead diver recorded fish data and a second diver followed behind, deploying the transect tape to ensure the area covered was undisturbed by diver presence (Emslie et al. 2018). In total, 14 species were recorded by UVC (Table 1). This included 9 of the 11 species recording >95% of total bites on *Xestospongia* spp. (see Table 2) and 4 additional pomacanthids (*Centropyge vroliki*, *Pomacanthus imperator*, *P. xanthometopon* and *P. sexstriatus*) that have been recorded grazing on sponges in other studies (Powell et al. 2015) and as part of an initial trial conducted in the present study (see Appendix).

2.5. Gut content analysis

To quantify the amount of sponge consumed by sponge-grazing fishes, 5 species (*C. binotatus*, *Chaetodon kleinii*, *Pygoplites diacanthus*, *Zanclus cornutus*, *Forcipiger flavissimus*) identified from *Xestospongia* spp. video footage as frequent sponge grazers were spear-fished at The Ridge and Sampela over 2 wk in July 2017 by local fishers (see Table 3 for sample sizes). Some species that were identified in video footage as frequent sponge grazers (e.g. *Coradion melanopus*) were not targeted due to low population size. We also captured one *P. imperator* from Sampela, and 3 *P. xanthometopon* and 5 *Siganus punctatus* from The Ridge; these species were not observed in videos of *Xestospongia* spp. but were observed feeding on other sponges in an initial trial (see Appendix). All fish were placed on ice immediately after spearing and kept at -20°C until dissection. During dissection, fish were weighed (precision: 0.1 g) and total length (mm) was recorded. The entire digestive tract was removed and preserved at -20°C in 96% ethanol to prevent further digestion. Preserved digestive tracts were returned to Wellington and dissected at the Victoria University Coastal Ecology Laboratory (VUCCEL). During dissection, contents from the stomach (or entire digestive tract for *C. binotatus* and *S. punctatus*) were examined under a dissecting microscope and separated into identifiable

Table 1. Sponge-grazing fish densities (mean ± SE) per 250 m² surveyed by underwater visual census (n = 6) at 4 sites in the Wakatobi Marine National Park

| Family / species | Sampela | B1 | The Ridge | KDS |
|----------------------------------|-----------|-----------|-----------|-----------|
| Acanthuridae | | | | |
| <i>Acanthurus pyroferus</i> | 4.8 ± 1.1 | 3.2 ± 0.4 | 5.0 ± 0.7 | 1.8 ± 0.7 |
| <i>Ctenochaetus binotatus</i> | 4.8 ± 1.0 | 4.3 ± 1.0 | 2.5 ± 0.5 | 3.3 ± 0.7 |
| Chaetodontidae | | | | |
| <i>Chaetodon kleinii</i> | 2.8 ± 0.8 | 4.3 ± 1.0 | 4.2 ± 1.5 | 2.8 ± 0.7 |
| <i>Coradion melanopus</i> | 0.0 ± 0.0 | 0.5 ± 0.5 | 0.0 ± 0.0 | 0.0 ± 0.0 |
| <i>Forcipiger flavissimus</i> | 1.3 ± 0.5 | 8.0 ± 0.9 | 5.5 ± 0.6 | 6.0 ± 1.4 |
| Pomacanthidae | | | | |
| <i>Centropyge bicolor</i> | 2.8 ± 0.9 | 0.0 ± 0.0 | 0.0 ± 0.0 | 0.3 ± 0.3 |
| <i>Centropyge nox</i> | 0.3 ± 0.2 | 0.0 ± 0.0 | 0.3 ± 0.3 | 0.0 ± 0.0 |
| <i>Centropyge tibicen</i> | 2.3 ± 0.9 | 0.5 ± 0.3 | 0.0 ± 0.0 | 0.0 ± 0.0 |
| <i>Centropyge vroliki</i> | 0.7 ± 0.4 | 0.0 ± 0.0 | 0.0 ± 0.0 | 0.3 ± 0.3 |
| <i>Pomacanthus imperator</i> | 0.0 ± 0.0 | 0.2 ± 0.2 | 0.0 ± 0.0 | 0.0 ± 0.0 |
| <i>Pomacanthus sexstriatus</i> | 0.0 ± 0.0 | 0.0 ± 0.0 | 0.0 ± 0.0 | 0.5 ± 0.5 |
| <i>Pomacanthus xanthometopon</i> | 0.2 ± 0.2 | 0.2 ± 0.2 | 0.0 ± 0.0 | 0.5 ± 0.3 |
| <i>Pygoplites diacanthus</i> | 0.2 ± 0.2 | 0.8 ± 0.4 | 2.7 ± 0.6 | 1.2 ± 0.4 |
| Zanclidae | | | | |
| <i>Zanclus cornutus</i> | 3.0 ± 0.5 | 2.8 ± 0.5 | 5.0 ± 0.8 | 4.8 ± 0.5 |

and unidentifiable portions. For *P. diacanthus*, *Pomacanthus* spp., *S. punctatus* and *Z. cornutus*, food items in the identifiable portion were separated into 13 major categories (see Table 3), blotted to remove gastric juices and weighed (wet weight: 0.01 g) to estimate diet proportions. Due to low stomach/digestive tract volumes, the diet composition of *C. kleinii*, *F. flavissimus* and *C. binotatus* was estimated visually via relative volumetric quantity (after Nielsen & Johnson 1992). Samples of sponges recovered from gut contents were stored in separate microcentrifuge tubes at -20°C in 96% ethanol for subsequent DNA extraction. All dissecting instruments were ethanol-cleaned and rinsed 3 times with double distilled water (DDW) between stomachs to prevent cross-contamination.

Levins (1968) standardised diet breadth was calculated as:

$$B = 1 / \sum p_i^2 \quad (1)$$

where B is niche breadth and p is the proportion of overall diet made up by the food category j . This can then be standardised and expressed on a scale of 0–1 according to the following:

$$BA = (B - 1) / (n - 1) \quad (2)$$

where BA is the standardised niche breadth and n is the number of food categories. Low values indicate a specialised diet based on few dietary items.

2.6. Sponge consumption estimates

To estimate sponge consumption, we used the mean weight (g) of sponge in the gut contents of each species and assumed that stomachs were filled at least twice daily following Harmelin-Vivien & Bouchoir-Navaro (1983). This weight was multiplied by the total number of fish of each species observed at each site and scaled to 1000 m^{-2} . All sponge consumption estimates are reported as mean \pm 95% CI.

2.7. DNA analysis of gut content sponges

Total genomic DNA was extracted from gut content sponges recovered from the angelfish *P. diacanthus*, *P. imperator*, and *P. xanthometopon* using a DNeasy Blood and Tissue Kit (Qiagen) following the protocol of the manufacturer. PCR was performed on extracted sponge DNA using custom 18S primers designed during this study (see Texts S1–S3 in the Supplement at www.int-res.com/articles/suppl/m672p123_supp.pdf for full details on primer design and trials) alongside a

sponge control, a predator control and a negative control. PCR was conducted in 25 μl reactions consisting of 12.5 μl MyTaq Red Mix (Bioline), 0.5 μl of each universal primer (10 μM), 0.5 μl of each blocking primer (100 μM), ~25 ng of template DNA and a volume of distilled water to reach 25 μl . An additional 5 μl of DMSO was added to the reaction to improve PCR yield. PCR cycling conditions were 94°C for 5 min; then 35 cycles of 94°C for 30 s, 48°C for 55 s, 72°C for 45 s and a final extension step of 72°C for 7 min. In total, 32 samples were successfully amplified and sent for purification and sequencing on an ABI 3730 capillary sequencer (Macrogen).

2.8. Data analysis

2.8.1. Video observations and fish densities

The difference in overall grazing intensity (count of bites; 35 species combined) between sites was tested using negative binomial regression using the MASS package (Venables & Ripley 2002) for R studio (RStudio Team 2015) with footage length (min) and area of sponge (m^2) included as offsets in the model. A difference in the density of sponge-grazing fishes recorded by UVC between sites was analysed using a 1-factor ANOVA, with test assumptions verified by Shapiro-Wilks test of normality and Levene's test of the equality of variances. Multivariate analyses were conducted with the PERMANOVA+ add-in to PRIMER v.6.1 (PRIMER-E). Differences in fish species contributing to grazing intensity (total bites) and in the community composition of sponge-grazing fishes between sites were assessed using Bray-Curtis similarity matrices with the unrestricted permutation of raw data and subsequent pair-wise comparisons. Data were square-root transformed prior to analysis and checked for multivariate homogeneity of group dispersions using PERMDISP. We used SIMPER (PRIMER-E) to calculate the contribution of each species (%) to the dissimilarity between sites.

2.8.2. Sequence data

Generated sequences were screened for contamination from other targets (e.g. fungus) and for duplicate samples (e.g. the same sponge sequence from the same stomach). Remaining sequences were quality checked using CodonCode Aligner (www.codoncode.com/aligner) by clipping bases until there were fewer than 3 bases with a phred quality score of

less than 20 in a 25 bp window and removing any sequences shorter than 200 bp in length. A demo-sponge data set was created using 18S rDNA sequence data downloaded from the GenBank database; the minimum inclusion was 500 bp and downloaded sequences were limited to the 6 orders represented by sequences obtained by this study. All sequences were aligned using the ClustalW function in BioEdit v.7.0.5.3 (Hall 1999). A phylogenetic tree was constructed of aligned sequence data using ModelFinder in IQ-TREE multicore v.1.6.11 (Kalyanamoorthy et al. 2017), which calculates the log-likelihoods of an initial parsimony tree for many different models, choosing the model that minimises the Bayesian information criterion (BIC). The glass sponge *Rhabdocalypus dawsoni* was used as an out-group (accession number AF100949) and branch supports were obtained using the SH-aLRT test (Guindon et al. 2010) and ultrafast bootstrap (Hoang et al. 2018). The best-fit model according to BIC was TIM2+F+I+G4. Branch supports are reported as SH-aLRT support (%) / ultrafast bootstrap support (%), and high support for the existence of a clade was determined by SH-aLRT $\geq 85\%$ / UFBoot $\geq 95\%$.

3. RESULTS

3.1. Video observations

In approximately 50 h of video footage, 28 477 bites were recorded on *Xestospongia* spp. across the 4 sites. In total, 33 species from 10 families of coral reef fish were identified grazing on *Xestospongia* spp. (Table 2). Grazing intensity (total bites) was significantly different between sites, with barrel sponges at Sampela experiencing fewer bites (mean \pm SE: 12.2 ± 2.6 bites $\text{min}^{-1} \text{m}^{-2}$) than barrel sponges at The Ridge (34.9 ± 10.8 bites $\text{min}^{-1} \text{m}^{-2}$) ($Z = 2.85$, $p = 0.004$) and barrel sponges at B1 (28.4 ± 6.0 bites $\text{min}^{-1} \text{m}^{-2}$) ($Z = 2.59$, $p = 0.009$) but not barrel sponges at KDS (21.7 ± 6.0 bites $\text{min}^{-1} \text{m}^{-2}$). Despite the large number of species observed grazing on *Xestospongia* spp., over half of the total bites recorded were taken by *Ctenochaetus* spp. (of the fish that could be identified to species, >95% of bites were attributed to *C. binotatus* and *Chaetodon kleinii*, and 95% were taken by 11 of the 33 species recorded (Table 2). Species contributing to grazing intensity were significantly different between sites (pseudo $F_{3,25} = 1.73$, $p = 0.012$), with pairwise tests indicating significant differences between B1 and Sampela ($p = 0.001$) and B1 and The Ridge ($p = 0.038$); all other pairwise comparisons

showed no significant differences ($p > 0.05$). The highest average within-group similarity was recorded at B1 (49.0%), followed by The Ridge (36.6%), KDS (34.8%) and Sampela (30.7%). SIMPER showed that the top 3 species driving the differences in predation between B1 and Sampela (average dissimilarity 68.7%), and B1 and The Ridge (average dissimilarity 63.4%) were *Coradion melanopus*, *Ctenochaetus* spp., and *C. kleinii*. At B1, *C. melanopus* took 18.8% of bites, in contrast to 1% at The Ridge and none at Sampela. At B1, *Ctenochaetus* spp. took 50.4% of bites, compared to 18.4% at Sampela and 19% at The Ridge. At B1, *C. kleinii* took 7.4% of bites, compared to 49.1% at Sampela and 33.7% at The Ridge (Fig. 2).

3.2. Sponge-grazing fish densities

Sponge-grazing fish densities (all species combined) were not statistically different between sites, with overall 24 ± 1 (mean \pm SE) fish observed per 250 m^2 . There was also no difference in the densities of the top 2 species observed grazing on *Xestospongia* spp. (Table 1). *C. binotatus* and *C. kleinii* (Table 1). However, there was a difference in sponge-grazing fish species composition between sites (pseudo $F_{3,20} = 3.95$, $p = 0.001$). Pairwise comparisons revealed that the sponge-grazing fish species composition at Sampela was significantly different from that at B1 ($t = 2.38$, $p = 0.003$), The Ridge ($t = 2.93$, $p = 0.003$) and KDS ($t = 2.19$, $p = 0.001$) and there was also a significant difference between The Ridge and B1 ($t = 1.55$, $p = 0.033$) (Fig. 3). SIMPER indicated that spatial variation in sponge-grazing fish species composition was mainly driven by variation in the abundance of *Forcipiger flavissimus*, *Centropyge bicolor* and *C. tibicen* (Table 1). At Sampela, densities of *C. bicolor* were 2.8 ± 0.9 fish per 250 m^2 but *C. bicolor* densities were only 0.3 ± 0.3 fish per 250 m^2 at KDS, and *C. bicolor* was not recorded at The Ridge or B1. Similarly, *C. tibicen* densities were 2.3 ± 0.9 per 250 m^2 at Sampela, but only 0.5 ± 0.3 per 250 m^2 at B1, and *C. tibicen* was not recorded at The Ridge or KDS. Densities of *F. flavissimus* were much lower at Sampela (1.3 ± 0.5 per 250 m^2) than at B1 (8 ± 0.9 per 250 m^2), The Ridge (5.5 ± 0.6 per 250 m^2) and KDS (6.0 ± 1.4 per 250 m^2).

3.3. Gut content analysis

Pygoplites diacanthus had stomach contents containing >90% sponge (% wet weight); it also had the

Table 2. Total number of bites taken by 33 reef fish species filmed grazing on *Xestospongia* spp. at 4 sites in the Wakatobi MNP. Frequency of occurrence in videos (FO) calculated from the total number of *Xestospongia* spp. pooled across all sites (n = 30). X: fish was observed biting *Xestospongia* spp. at a particular site; (*) species sampled for gut contents

| Species | Family | Total bites | Cumulative (%) | FO (%) | Sam-pela | B1 | The Ridge | KDS |
|--|----------------|-------------|----------------|--------|----------|----|-----------|-----|
| <i>Ctenochaetus</i> spp.* | Acanthuridae | 10771 | 37.8 | 80 | X | X | X | X |
| <i>Chaetodon kleinii</i> * | Chaetodontidae | 7829 | 65.3 | 86.7 | X | X | X | X |
| <i>Zanclus cornutus</i> * | Zanclidae | 1835 | 71.8 | 36.7 | X | X | X | X |
| <i>Coradion melanopus</i> | Chaetodontidae | 1687 | 77.7 | 26.7 | | X | X | X |
| <i>Forcipiger flavissimus</i> * | Chaetodontidae | 1545 | 83.1 | 40 | | X | X | X |
| <i>Centropyge tibicen</i> | Pomacanthidae | 1167 | 87.2 | 13.3 | X | X | | X |
| <i>Scarus flavipectoralis</i> juv. | Scaridae | 996 | 90.7 | 26.7 | X | X | | X |
| <i>Pygoplites diacanthus</i> * | Pomacanthidae | 595 | 92.8 | 20 | X | X | X | X |
| <i>Heniochus singularis</i> | Chaetodontidae | 282 | 93.8 | 16.7 | | X | X | |
| <i>Centropyge bicolor</i> | Pomacanthidae | 259 | 94.7 | 10 | X | | | X |
| <i>Centropyge nox</i> | Pomacanthidae | 225 | 95.5 | 10 | X | X | X | |
| <i>Zebrasoma scopas</i> | Acanthuridae | 163 | 96.1 | 13.3 | | X | X | X |
| <i>Thalassoma lunare</i> | Labridae | 156 | 96.6 | 23.3 | X | X | | X |
| <i>Chaetodon punctatofasciatus</i> | Chaetodontidae | 138 | 97.1 | 10 | | X | | X |
| <i>Balistapus undulatus</i> | Balistidae | 105 | 97.5 | 33.3 | X | X | X | X |
| <i>Centropyge vrohki</i> | Pomacanthidae | 102 | 98.2 | 10 | X | | | X |
| <i>Chromis xanthochira</i> | Pomacentridae | 102 | 97.8 | 10 | X | | X | |
| <i>Anampses meleagrides</i> | Labridae | 89 | 98.5 | 13.3 | X | X | | X |
| <i>Scarus niger</i> | Scaridae | 83 | 98.8 | 13.3 | | X | X | X |
| <i>Halichoeres prosopion</i> | Labridae | 69 | 99.0 | 26.7 | | X | X | X |
| <i>Neoglyphidodon melas</i> | Pomacentridae | 65 | 99.2 | 13.3 | | X | X | X |
| <i>Diproctacanthus xanthurus</i> | Labridae | 51 | 99.4 | 6.7 | | X | X | |
| <i>Canthigaster valentini</i> | Tetraodontidae | 38 | 99.6 | 3.3 | | | | X |
| <i>Heniochus pleurotaenia</i> | Chaetodontidae | 37 | 99.7 | 3.3 | | X | | |
| <i>Scarus dimidiatus</i> | Scaridae | 22 | 99.8 | 6.7 | X | X | | |
| <i>Pomacentrus adelus</i> | Pomacentridae | 13 | 99.8 | 3.3 | X | | | |
| <i>Scarus ghobban</i> | Scaridae | 11 | 99.9 | 3.3 | | | | X |
| <i>Paracentropyge multifasciata</i> | Pomacanthidae | 10 | 99.9 | 3.3 | | X | | |
| <i>Siganus vulpinus</i> | Siganidae | 9 | 99.9 | 3.3 | | | | X |
| <i>Chaetodon ulietensis</i> | Chaetodontidae | 8 | 99.9 | 3.3 | | | X | |
| <i>Sufflamen bursa</i> | Balistidae | 6 | 99.9 | 10 | X | | X | |
| <i>Melichthys vidua</i> | Balistidae | 6 | 99.9 | 3.3 | | | | X |
| <i>Odonus niger</i> | Balistidae | 3 | 100.0 | 3.3 | | X | | |
| Other species grazing on sponges where bites could not be accurately determined: | | | | | | | | |
| <i>Acanthurus pyroterus</i> | Acanthuridae | | | 63.7 | X | X | X | X |
| <i>Acanthurus auranticavus</i> | Acanthuridae | | | 10 | X | | | |
| <i>Acanthurus nigricauda</i> | Acanthuridae | | | 3.3 | X | | | |
| <i>Acanthurus thompsoni</i> | Acanthuridae | | | 3.3 | X | | | |

narrowest diet breadth of all sampled species (diet breadth = 0.005) (Table 3). Soft coral was found in the stomach of one individual, and small amounts of tunicates and mobile invertebrates including polychaetes and small bivalves were also observed. The one *Pomacanthus imperator* stomach examined contained >90% sponge (% wet weight).

Sponge made up approximately half of the gut contents (% wet weight) of *Pomacanthus xanthometopon* stomachs, but soft coral (genus *Dendronephthya*) occurred in equal quantities (Table 3). Less frequent prey items included gastropods, polychaetes, and one stomach contained several juvenile starfish. Sponges were also found to be a major prey item of

Zanclus cornutus, making up approximately 70% of total weighed contents. Other items observed included macroalgae and calcareous sediments. *Siganus punctatus* contained approximately one-third sponge (% weight), with the most dominant dietary item identified as the red alga *Laurencia* spp. Seagrass was present in all *S. punctatus* stomachs examined, as were other types of algae and small mobile invertebrates were occasionally observed.

C. kleinii had one of the widest diets observed of all the fish sampled for gut contents (Table 3). Although video observations of *Xestospongia* spp. suggested this species was spongivorous, sponge tissue was not observed, although sponge spicules were

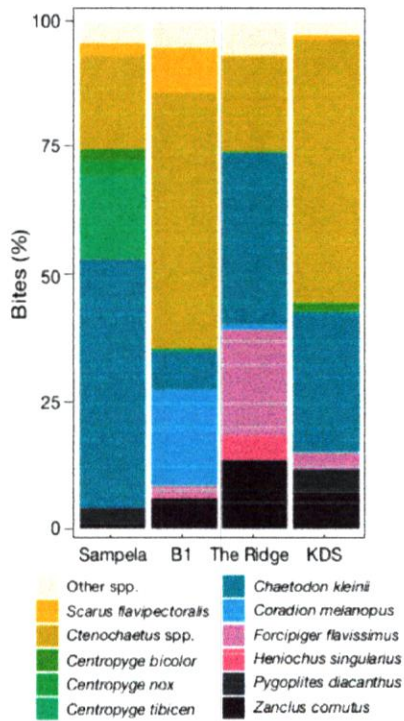


Fig. 2. Percentage of total bites on *Xestospongia* spp. taken by 11 species of sponge-grazing fishes filmed at 4 sites in the Wakatobi MNP

seen in 9 of 12 stomachs. Most of the organic matter in *C. kleinii* appeared to be of soft coral origin, and a variety of crustacea were observed. Small volumes of green filamentous algae were also present in most stomachs. *F. flavissimus* appeared to have a much more specialised diet than *C. kleinii*, with the majority of stomachs containing tentacles of *Serpulidae* worms (possibly *Spirobranchus* as distinctive toothed opercula were visible in many samples). Sponge tissue remains were sighted in 2 *F. flavissimus* stomachs, visually estimated to be <5% of the total volume. Other minor prey items for *F. flavissimus* included amphipods, isopods and decapod shrimps. The gut contents of *C. binotatus* were dominated by calcareous sediments and unidentifiable organic material. Sponge spicules were sighted in 9 of 12 gut contents but in very low numbers (1 or 2 gut⁻¹).

3.4. Sponge consumption

Estimates of sponge consumption were highest at the 2 outer reef sites, with 73.3 ± 21.5 g sponge

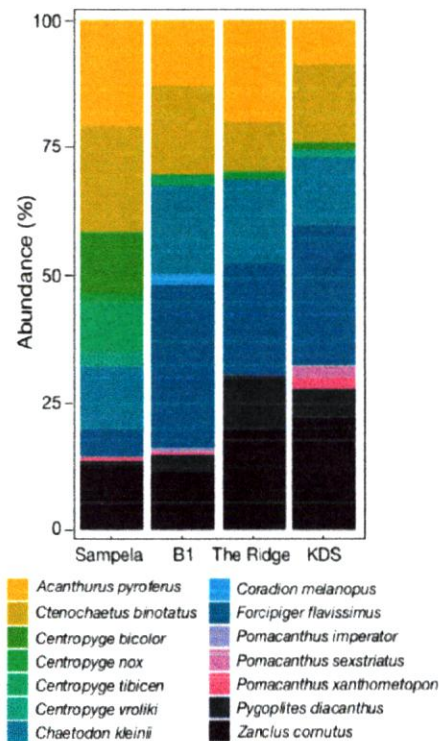


Fig. 3. Percentage of total abundance of sponge-grazing fishes surveyed by underwater visual census at each of the 4 study sites in the Wakatobi MNP

$1000 \text{ m}^{-2} \text{ d}^{-1}$ (mean \pm 95% CI) consumed at The Ridge and 52.9 ± 25.6 g sponge $1000 \text{ m}^{-2} \text{ d}^{-1}$ consumed at KDS (Table 4). The lowest estimate was at Sampela, with 27.6 ± 15.3 g sponge $1000 \text{ m}^{-2} \text{ d}^{-1}$; 32.5 ± 10.9 g sponge $1000 \text{ m}^{-2} \text{ d}^{-1}$ was consumed at B1. Across all reefs, the 5 spongivorous fishes identified by this study consumed 46.6 ± 18.3 g sponge $1000 \text{ m}^{-2} \text{ d}^{-1}$.

3.5. DNA analysis of consumed sponges

Out of 32 samples sent for sequencing, 7 amplified fungus, 2 amplified crinoid and 23 amplified sponge. Of the 23 sponge sequences, only 14 met the stringent quality-control standards employed by this study. The identity of one of these samples could not be resolved beyond Porifera and was removed from the data set. The remaining 13 sponges represented 6 orders of Porifera (Fig. 4). Six sponges were placed into Order Haplosclerida and came from *P. diacanthus* specimens collected at both Sampela and The Ridge (Fig. S2). The 18S

data provided strong support that one sponge (Ridge RA2 S10) was placed into a clade with the Indo-Pacific sponge *Chalinula hooperi* (Bakus & Nishiyama 2000) (91.7/100) and another sponge (Ridge RA5) had strong support for placement within a clade of *Petrosia* sponges (98.6/96). The sponge Ridge RA1 S4A also formed a highly supported clade (97.8/99) with the species *Haliclona*

curacaoensis and *Niphates erecta*. For the remaining samples in this order, identities could not be resolved beyond Haplosclerida. The Haplosclerida sponges formed 4 weakly supported major clades. All of the sponges placed in Order Haplosclerida were placed within a weakly supported major clade containing mainly Petrosiidae and Niphatidae sponges, with some interspersed *Haliclona*.

Table 3. Relative gut content (%) of 8 species of sponge-grazing reef fish caught in the Wakatobi MNP in 2017. *P. diacanthus* to *S. punctatus* are given in % wet weight (%W) and *C. kleini* to *C. binotatus* are relative volumetric quantity (% volume, %V) due to low volume of stomach/gut contents. Frequency of occurrence (%F) of food items indicated. Bold values indicate sponge content >20%. TL: total length

| | <i>Pygoplites diacanthus</i> | | <i>Pomacanthus imperator</i> | | <i>Pomacanthus xanthurus</i> | | <i>Zanclus cornutus</i> | | <i>Siganus punctatus</i> | | <i>Chaetodon kleinii</i> | | <i>Forcipiger flavissimus</i> | | <i>Ctenochaetus binotatus</i> | |
|----------------------|------------------------------|-----|------------------------------|----|------------------------------|-----|-------------------------|-----|--------------------------|-----|--------------------------|-----|-------------------------------|-----|-------------------------------|----|
| Gut samples (n) | 13 | | 1 | | 3 | | 15 | | 5 | | 11 | | 11 | | 10 | |
| TL range (mm) | 122–188 | | 102 | | 204–272 | | 82–150 | | 236–285 | | 82–104 | | 122–158 | | 122–136 | |
| Levins' diet breadth | 0.005 | | 0.004 | | 0.046 | | 0.036 | | 0.095 | | 0.097 | | 0.032 | | N/A | |
| | %W | %F | %W | %F | %W | %F | %W | %F | %W | %F | %V | %F | %V | %F | %V | %F |
| Annelida | | | | | | | | | | | | | | | | |
| Polychaete | <1 | 69 | <1 | | <1 | 33 | 1 | 53 | <1 | 60 | 15 | 92 | 75 | 100 | | |
| Crustacea | | | | | | | | | | | | | | | | |
| Amphipod | | | <1 | | | | <1 | 13 | <1 | 60 | 14 | 58 | 7 | 27 | | |
| Copepod | | | | | | | | | | 20 | 4 | 50 | | | | |
| Decapod | | | | | | | <1 | 7 | <1 | 20 | <1 | 17 | 9 | 18 | | |
| Isopod | | | | | | | <1 | 7 | | | 4 | 33 | 2 | 9 | | |
| Bryozoa | | | | | | | | | | | | | | | | |
| Bryozoan | | | | | | | 1 | 20 | | | | | | | | |
| Porifera | | | | | | | | | | | | | | | | |
| Sponge | 94 | 100 | 96 | | 53 | 100 | 73 | 100 | 32 | 100 | <1 | 75 | 2 | 18 | <1 | 70 |
| Chordata | | | | | | | | | | | | | | | | |
| Ascidian | <1 | 38 | 1 | | <1 | 33 | 3 | 73 | 17 | 100 | | | | | | |
| Echinodermata | | | | | | | | | | | | | | | | |
| Asteroidea | | | | | <1 | 33 | | | | | | | | | | |
| Ophiuroidea | | | | | | | | | <1 | 20 | | | | | | |
| Cnidaria | | | | | | | | | | | | | | | | |
| Hydrozoa | | | | | | | | | | | 2 | 42 | | | | |
| Anthozoa—hard coral | | | | | <1 | 100 | <1 | 13 | | | 2 | 33 | | | | |
| Anthozoa—soft coral | 5 | 8 | 3 | | 46 | 100 | 10 | 7 | <1 | 20 | 51 | 83 | <1 | 9 | <1 | 50 |
| Mollusc | | | | | | | | | | | | | | | | |
| Gastropod | | | | | <1 | 33 | <1 | 20 | | | <1 | 17 | | | | |
| Bivalve | <1 | 46 | | | | | 1 | 60 | <1 | 60 | <1 | 17 | 1 | 18 | | |
| Algae | | | | | | | | | | | | | | | | |
| Calcareous algae | | | | | <1 | 33 | 1 | 73 | 1 | 20 | | | | | | |
| Red macroalgae | | | | | | | | | | | 42 | 100 | | | | |
| Green filamentous | <1 | 46 | | | | | | | | | 4 | 58 | | | 5 | 80 |
| Green turfing | | | | | | | | | <1 | 20 | | | | | | |
| Brown macroalgae | | | | | | | 2 | 60 | <1 | 40 | | | | | | |
| Seagrass | | | | | | | | | | | | | | | | |
| Seagrass | | | | | | | | | 6 | 80 | | | | | 4 | 90 |
| Eggs | | | | | | | | | | | | | | | | |
| Eggs | | | | | <1 | 6 | | | | | 3 | 50 | <1 | 9 | | |
| Sediments | | | | | | | | | | | | | | | | |
| Calcareous deposits | <1 | 15 | | | | | 6 | 67 | 1 | 60 | | | | | | 90 |
| Fluid | | | | | | | | | | | | | | | | |
| Fluid/mucus | | | | | | | | | | | | | 2 | 9 | | |

Table 4 Estimates of sponge consumption for 5 species of spongivorous reef fish calculated from mean sponge weight in gut contents and fish densities. NB: *Siganus punctatus* was not included in the underwater visual census of sponge-grazing fishes, therefore, density was conservatively estimated from video observations

| Site / spongivore | Total fish in 1500 m ² | Sponge weight (g) (mean ± 95% CI) | Sponge consumption (mean ± 95% CI) (g 1000 m ⁻² d ⁻¹) |
|----------------------------------|-----------------------------------|-----------------------------------|--|
| Sampela | | | |
| <i>Pygoplites diacanthus</i> | 1 | 2.05 ± 0.38 | 2.7 ± 0.5 |
| <i>Pomacanthus imperator</i> | 0 | 2.83 | 0.0 ± 0.0 |
| <i>Pomacanthus xanthometopon</i> | 1 | 4.68 ± 4.54 | 6.2 ± 6.1 |
| <i>Zanclus cornutus</i> | 18 | 0.39 ± 0.10 | 9.4 ± 2.4 |
| <i>Siganus punctatus</i> | 2 | 3.49 ± 2.35 | 9.3 ± 6.3 |
| B1 | | | |
| <i>P. diacanthus</i> | 5 | 2.05 ± 0.38 | 13.7 ± 2.5 |
| <i>P. imperator</i> | 1 | 2.83 | 3.8 |
| <i>P. xanthometopon</i> | 1 | 4.68 ± 4.54 | 6.2 ± 6.1 |
| <i>Z. cornutus</i> | 17 | 0.39 ± 0.10 | 8.8 ± 2.3 |
| <i>S. punctatus</i> | 0 | 3.49 ± 2.35 | 0.0 ± 0.0 |
| The Ridge | | | |
| <i>P. diacanthus</i> | 16 | 2.05 ± 0.38 | 43.7 ± 8.1 |
| <i>P. imperator</i> | 0 | 2.83 | 0.0 ± 0.0 |
| <i>P. xanthometopon</i> | 0 | 4.68 ± 4.54 | 0.0 ± 0.0 |
| <i>Z. cornutus</i> | 30 | 0.39 ± 0.10 | 15.6 ± 4.0 |
| <i>S. punctatus</i> | 3 | 3.49 ± 2.35 | 14.0 ± 9.4 |
| KDS | | | |
| <i>P. diacanthus</i> | 7 | 2.05 ± 0.38 | 19.1 ± 3.5 |
| <i>P. imperator</i> | 0 | 2.83 | 0.0 ± 0.0 |
| <i>P. xanthometopon</i> | 3 | 4.68 ± 4.54 | 18.7 ± 18.2 |
| <i>Z. cornutus</i> | 29 | 0.39 ± 0.10 | 15.1 ± 3.9 |
| <i>S. punctatus</i> | 0 | 3.49 ± 2.35 | 0.0 ± 0.0 |

The second order that was represented by more than one sponge was Verongiida, and all of these sponges came from stomachs of *P. diacanthus* collected at Sampela (Fig. S3). There was strong support for all 3 sponges being placed within a clade containing Pseudoceratinidae and *Verongula* sponges (86.5/97). The remaining sponges were all classified as separate orders and were sequenced from the stomachs of *P. diacanthus* and *P. xanthometopon* collected at The Ridge. There was strong support for the placement of one sponge (Ridge RA6 S3) within a clade of Agelasidae sponges (87.5/95) and another sponge (Ridge RA4 B) had strong support for placement within a clade of *Aaptos* sponges from the order Suberitida (98.7/100). The sponge Ridge RA3 S1 had strong support for placement within a clade containing both Clionaidae and Spirastrellidae genera (97.2/99) and the only sponge that was successfully sequenced from the gut contents of *P. xanthometopon* (Ridge PX2 S13) was grouped in Order Poecilosclerida and had strong support for placement into a clade containing *Guitarra* sp. (88.1/100).

4. DISCUSSION

Resolving predator–prey dynamics is essential to understand the structure of ecological communities (Leray et al. 2015, Nielsen et al. 2018). Our study provides the first estimate of sponge biomass consumption on coral reefs and is the first to successfully sequence sponges recovered from the stomach contents of spongivorous fishes. Five important spongivores were identified through a combination of video observations and gut content analysis and were estimated to consume 46.6 ± 18.3 g sponge 1000 m⁻² d⁻¹ (mean ± 95% CI). Interestingly, the most prolific sponge-grazers (*Chaetodon kleinii* and *Ctenochaetus* spp.) had no sponge tissue remains in gut contents, highlighting the importance of using direct methods of diet quantification to prevent functional misidentifications (see Bellwood et al. 2019). Fourteen unique sponge sequences were amplified from sponges found in angelfish stomach contents, suggesting that *Pygoplites diacanthus* exhibited high dietary plasticity within Phylum Porifera similar to the diets of angelfishes in the Caribbean (Randall & Hartman 1968) and the Pacific (Verdín Padilla et al. 2010).

4.1. Identified spongivores

Angelfishes are well-known spongivores in the Caribbean and the Pacific (Randall & Hartman 1968, Hobson 1974, Verdín Padilla et al. 2010), with at least 5 species thought to rely on sponges for the majority of their diet (Randall & Hartman 1968, Hobson 1974). The present study confirms that *P. diacanthus* may also be reliant on sponges as stomach contents contained >80% sponge. Our findings are consistent with previous *in situ* observations of *P. diacanthus* feeding behaviour (Powell et al. 2015). *P. diacanthus* had the largest predatory impact on sponges, accounting for almost half of the total sponge consumption. The one *Pomacanthus imperator* stomach dissected contained >90% sponge, suggesting a similar reliance on sponges to *P. diacanthus*, but more samples would be required to determine if *P. imperator* is a true obligate spongivore on Wakatobi reefs. The *P. imperator*

stomach examined was dominated by sponges that appeared to be highly spiculate, whereas *P. diacanthus* stomachs contained both highly silicified sponges and fewer but noticeable aspiculate sponges with organic fibres present. Gut content data also confirmed *Pomacanthus xanthometopon* was a spongivore, but it appeared to have a lower reliance on sponges than *P. diacanthus*, consuming an equal amount of predominantly *Dendronephthya* soft corals, with one stomach containing 3 juvenile starfish.

Sponges were the dominant dietary item for *Zanclus cornutus*, although its diet also consisted of algae and other sessile invertebrates, including bryozoans and ascidians. The rabbitfish *Siganus punctatus* also regularly consumed sponges, which were found in the gut contents of all sampled individuals, although in widely varying amounts (2–76% wet weight, g). Our results agree with data from the Great Barrier Reef and Japan, where sponges represented 20 and 72% of the gut contents of adult *S. punctatus* (Pitt 1997, Sano 1989). Our data indicate that sponges and the red alga *Laurencia* spp. made up the majority of the diet of adult *S. punctatus* on Wakatobi reefs. Seagrass was also present in small amounts (6% wet weight, g) in 4 out of 5 individuals. This may have been ingested in reef areas as detritus, but could also indicate movement between intertidal seagrass beds and shallow reef areas.

It is a commonly held belief that spongivorous angelfish feed rotationally in a 'smorgasbord' fashion (Randall & Hartman 1968), perhaps to reduce the impact of secondary metabolites (Wulff 2006). Video observations in the present study agree with observations noted in Pawlik et al. (2018) that spongivores can demonstrate repetitive and focused feeding on palatable sponges. *P. imperator*, *P. diacanthus* and *Z. cornutus* were recorded taking numerous bites from the exposed choanosome of *Petrosia corticata* (Fig. A1), with many successive bites taken in single feeding events (e.g. 34 bites for *P. imperator*). Our observations also suggest that the cyanobacterial ectosome of *P. corticata* may have been conferring some degree of anti-predator defence as it was avoided by the fish, which took repetitive bites from the inner exposed tissue.

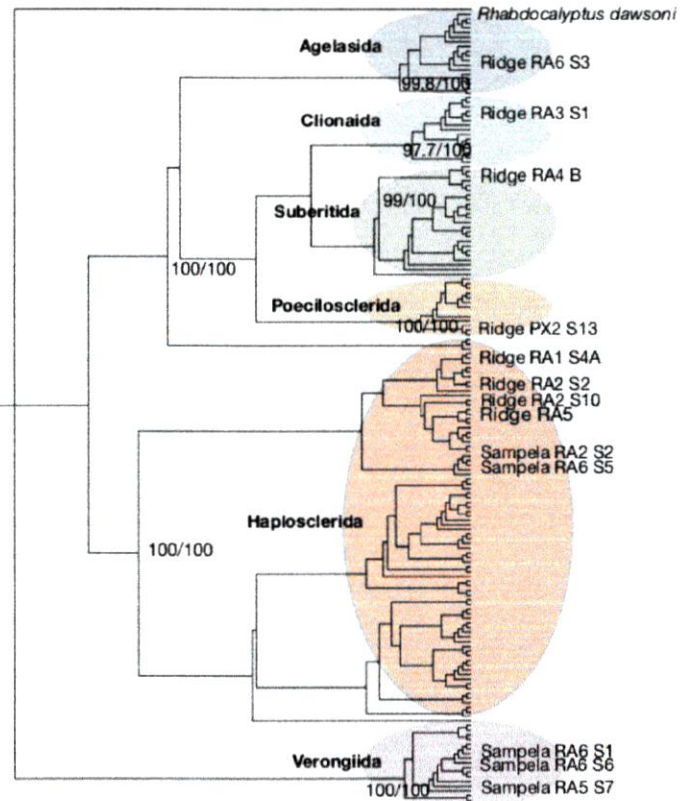


Fig. 4. Cladogram showing the placement of 13 sponges within 6 orders of Porifera (shown by colored ellipses) dissected from the stomach contents of the angelfishes *Pygoplites diacanthus* (RA) and *Pomacanthus xanthometopon* (PX) caught at 2 sites in the Wakatobi MNP and based on 18S sequence data. The outgroup is the glass sponge *Rhabdocalyptus dawsoni* AF100949). Branch supports reported as SH-aLRT support (%) / ultrafast bootstrap support (%)

4.2. Key sponge-grazing fishes not consuming sponge tissue

Despite recording >7500 bites on *Xestospongia* spp. sponge tissue, remains were not found in *C. kleinii* stomach contents, although spicules were sighted in 9 of 12 stomachs. Similarly, *Forcipiger flavissimus* recorded >1500 bites on *Xestospongia* spp. but out of 11 stomachs analysed, only 2 contained sponge tissue remains, both amounting to 2% of the total stomach content. Our findings indicate that for some fishes (e.g. butterflyfishes), bites on sponge surfaces do not provide reliable evidence of sponge consumption. Angelkerken et al. (2009) made similar observations on a fringing reef in East Kalimantan, Indonesia, where *C. kleinii* was observed taking ap-

proximately one-third of bites from sponges but stomach contents contained only small volumes of sponge (<8% by volume of identifiable contents). The small volumes of sponge in the gut contents of chaetodontids detailed in this study and other studies from the Indo-Pacific (Sano 1989, Nagelkerken et al. 2009) suggest that butterflyfishes are likely targeting sponge-associated macrofauna. Spongivorous polychaetes (e.g. *Haplosyllis spongicola* and *Branchiosyllis oculata*) can directly penetrate sponge tissues where canals are small (Martin & Britayev 1998) and have been found in extremely high densities on both the surface and inside sponges (Pawlik 1983). Observations of the internal tissue of *Xestospongia* spp. in the Wakatobi MNP have found very high densities of an unidentified white polychaete (Family Syllidae) (579 ± 44 ind. g^{-1} [dry weight]) (C. Mortimer unpubl. data), which may explain the high grazing rates by chaetodontids and other invertivores.

Ctenochaetus binotatus likely lacks the necessary dental morphology to remove sponge biomass, instead appearing to 'brush' the surface of the sponge with comb-like teeth (Purcell & Bellwood 1993). Members of this genus are generally considered detritivores (Choat et al. 2002); however, other studies have indicated that *C. striatus* can also remove large quantities of algal turf (Marshall & Mumby 2015). Despite numerous observations of *Xestospongia* spp. grazing, gut content data presented here supports the conclusion that *C. binotatus* in the Wakatobi MNP is a detritivore that has no predatory impact on sponges. Sponge-grazing by detritivorous fishes may relate to the presence of mucus-bound sediment as some sponges, including *Xestospongia* spp., produce mucus as a sediment-clearing mechanism (McGrath et al. 2017). The sponge-associated holothuroid *Synaptula lamperti* ingests mucus directly from the sponge *lanthella basta*, incorporating metabolised compounds into body wall tissues (Hammond & Wilkinson 1985). Little is known about sponge mucus; however, coral mucus is known to be energy- and nutrient-rich (Wild et al. 2004) and is consumed by some reef fishes (Cole et al. 2010). The feeding behaviour of *C. binotatus* could also be explained by the 'sponge loop', whereby sponges take up dissolved organic carbon, which is converted into cellular detritus and readily ingested by sponge-associated invertebrates (de Goeij et al. 2013, Rix et al. 2018). However, the ubiquity of the sponge loop has rarely been proven beyond observations of cryptic, encrusting sponges (but see Rix et al. 2018), and the extent to which the sponge loop supports higher trophic levels across global reef ecosystems is currently unknown (Bell et al. 2018b).

4.3. Sponge consumption

Quantitative data on the role of sponges in the diets of spongivorous fishes combined with density data made it possible to estimate sponge consumption at each location. Considering these are the first known estimates of sponge consumption, comparisons with studies from other regions are not possible. However, following the same methodology as Harmelin-Vivien & Bouchon-Navaro (1983) allows for a comparison with coral consumption in the Central Pacific. The maximum coral consumption reported was 62.53 coral $1000 m^{-2} d^{-1}$ at the outer barrier reef slope (Harmelin-Vivien & Bouchon-Navaro 1983), which is similar to the estimates of daily sponge consumption reported for the outer reef sites in this study (mean: 52.9 – 73.3 g sponge $1000 m^{-2} d^{-1}$), suggesting that the biomass transferred from sponges to spongivores may be of a similar magnitude to the coral to corallivore pathway, although this comparison is limited by large geographic differences. Nevertheless, data presented here suggests that spongivory may be an important but relatively overlooked trophic link on Wakatobi reefs.

Interestingly, we found that sponge consumption was lowest on the sponge-dominated reef site of Sampela (27.6 ± 15.3 g sponge $1000 m^{-2} d^{-1}$), mainly driven by low densities of the 5 identified spongivorous fishes. It has been suggested that increasing sponge dominance may support a greater abundance of spongivorous fishes (Bell et al. 2013), and Powell et al. (2014) recorded a higher density of spongivorous fishes at Sampela relative to surrounding reefs (Powell 2013, Powell et al. 2014). Our results likely differ from Powell et al. (2014) due to the differences in the species categorised as spongivores. We used gut content analysis to confirm 'true' spongivores (i.e. those removing sponge biomass), whereas Powell et al. (2014) used *in situ* observations to categorise several species as spongivores that we suggest have no quantifiable predatory impact on sponge assemblages (e.g. *C. binotatus*). Results from the present study suggest that spongivores may not necessarily benefit from shifts to sponge dominance that are driven by a decline in corals involve the increase of 1 or 2 competitive sponge species (e.g. Aronson et al. 2002, Powell et al. 2014). *Lamellodysidea herbacea* accounts for approximately 40% of total sponge abundance at Sampela based on area occupied (Powell et al. 2014, Biggerstaff et al. 2015) but does not appear to be preferred or readily consumed by spongivores. The only spongivorous fish filmed taking bites from *L. herbacea* was *Z. cornutus* (see the

Appendix), although tissue removal could not be verified by video analysis. Additionally, DNA analysis did not identify *L. herbacea* in the gut contents of spongivorous angelfishes. *L. herbacea* has abundant cyanobacterial symbionts that produce polybrominated diphenyl ethers that may function as anti-predator compounds (Becerro et al. 2003), which could be another factor leading to the proliferation of *L. herbacea* at Sampela.

4.4. Taxonomy of consumed sponges

From 13 angelfish stomachs, 14 unique sponge sequences were retrieved spanning 6 orders of Porifera, suggesting substantial dietary plasticity, similar to the diets of angelfish in the Caribbean (Randall & Hartman 1968) and the Pacific (Verdin Padilla et al. 2010). The majority of sequences were placed into Order Haplosclerida, an extremely diverse order of Porifera (van Soest & Hooper 2002). There was strong support for the placement of one sponge into a clade with the Indo-Pacific sponge *Chalinula hooperi* (Bakus & Nishiyama 2000), which is commonly found in the shallow waters of pristine, offshore reefs (Cleary & de Voogd 2007). There was also strong support for the placement of one sponge into a clade containing sponges from the genus *Petrosia*, and another sponge had high support for placement alongside 2 Caribbean sponges *Niphates erecta* and *Haliclona curacaoensis*. *Haliclona* spp. have been reported from the gut contents of spongivorous fishes in the Caribbean (Randall & Hartman 1968, Hourigan et al. 1989) and the Pacific (Verdin Padilla et al. 2010). Similarly, *N. erecta* makes up a substantial proportion of the diet of Caribbean angelfishes as it is undefended by secondary metabolites or structural components and invests heavily in growth to compensate for biomass removal (Pawlik 2011).

Three sponges in the Verongiida group had high support for placement into a clade alongside sponges from the genera *Pseudoceratina* and *Verongula*. Verongiida sponges lack spicules; instead, they are structurally supported by 3-dimensional internal scaffolds made of fibrous chitin (Zóltowska-Aksamitowska et al. 2018), which may lessen their digestibility and cause them to persist in stomach contents at the expense of sponges with more diffuse tissue. Whilst this is quite possible, an unknown species of *Verongula* was also observed in the stomach contents of *Holacanthus tricolor* in the US Virgin Islands (Hourigan et al. 1989), and Verongiida sponges appear to form some part of the diet of *P. diacanthus*

as these sponges were amplified by 2 fish caught at Sampela reef. Of the sequences amplified from fish caught at Sampela reef, 3 were Verongiida sponges and 2 were placed in Order Haplosclerida, whereas at the outer reef site, sequences were amplified from 5 orders of Porifera. This could be indicative of differences in sponge assemblage composition, which can vary substantially over short spatial scales in the Wakatobi MNP (Powell et al. 2014, Mortimer 2020). It may also reflect differences in diversity, as Sampela reef has substantially lower sponge diversity than The Ridge (Powell et al. 2014, Mortimer 2020). There was strong support for one sponge belonging to a clade with 2 species of *Aaptos*. *A. subretoides* is commonly found on Wakatobi reefs and has been found to produce numerous compounds with antitumor, antimicrobial and antiviral activity (reviewed in Cleary et al. 2018). Meylan (1988) documented the closely related *A. aaptos* as one of the 10 most important sponges eaten by the hawksbill turtle, even though it produces toxic secondary metabolites (Nakamura et al. 1982). There was also strong support for the placement of one sponge into a clade of sponges from the family Agelasidae that are known for producing bromopyrrole alkaloids (Wilson et al. 1999, Lindel et al. 2000) that deter predation (Pawlik et al. 1995). An unknown species of *Agelas* represented the largest volume of sponge in the gut contents of the angelfish *Holacanthus ciliaris* and was also consumed by a filefish (Randall & Hartman 1968). The presence of sponges with substantial chemical defence is interesting because angelfishes are thought to mainly consume chemically undefended, palatable sponges (Pawlik 2011, Pawlik et al. 2018). However, without robust dietary proportion data, the extent of predation on particular Indo-Pacific sponges by angelfishes remains unknown.

There was strong support for one sponge belonging to a clade containing the bio-eroding families Clionaidae and Spirastrellidae. This is consistent with observations from the Caribbean which have recorded bio-eroding sponges from the genera *Spirastrella*, *Cliona* and *Spechiospongia* in angelfish gut contents (Randall & Hartman 1968). The only sponge that could be amplified from the gut contents of *P. xanthometopon* had strong support for placement into a clade of sponges from the genus *Guitarra* (Carter 1874) from the Guitarridae family. To date, there are 15 described species of *Guitarra*, but none have a documented distribution in the Indo-Pacific (<https://www.marinespecies.org>). Finally, the identity of one sample could not be resolved beyond the

higher classification of Porifera. This is likely due to poor data coverage for sponges from the Indo-Pacific, one of the most speciose regions of Porifera on the planet (van Soest et al. 2012).

4.5. Limitations of sponge consumption estimates

Sponge consumption estimates provided were conservatively based on stomach contents rather than total gut contents, and on the assumption that the spongivores identified (*P. diacanthus*, *P. imperator*, *P. xanthometopon*, *Z. cornutus* and *S. punctatus*) filled their stomachs twice daily. Coral material has been estimated to pass through the intestinal tract of the butterflyfish *Chaetodon unimaculatus* in 1.5–2 h, indicating that stomachs may be filled at least 6 times daily (Cox 1986, Cole et al. 2011). Considering that reported gut length to body length ratios for *C. unimaculatus* (Berumen et al. 2011) and *Pomacanthus* spp. (Pérez-España & Abitia-Cárdenas 1995) are similar, it seems likely that angelfishes may also fill their stomachs up to 6 times daily. However, we chose to follow the same methods as Harmelin-Vivien & Bouchon-Navaro (1983) to facilitate a comparison with coral consumption. Other methods of estimating consumption were considered, such as the method employed by Cole et al. (2011); however, retaining fish in aquaria to conduct feeding trials was not feasible at this study site.

This study was limited to vertebrate sponge predators that are thought to have the biggest predatory impact on sponge assemblages on shallow coral reefs (Randall & Hartman 1968, Meylan 1988). However, tropical sponges also have a range of invertebrate predators including echinoderms, molluscs, crustaceans and annelids (see Bell et al. 2020 for review), although the effect of invertebrate predation on tropical sponge populations and species diversity is largely unknown. Additionally, other spongivorous vertebrates considered to exhibit top-down control on sponge populations in the Caribbean were not considered in our estimates of sponge consumption. Hawksbill turtles are found in extremely low abundances on Wakatobi reefs (only one turtle was sighted during the 3 mo study period), hence their impact on sponge assemblages in the Wakatobi MNP is likely negligible. Nevertheless, before populations were decimated by over-exploitation, they likely had a strong structural effect on sponge populations as they are known to feed almost exclusively on a narrow range of available sponges (Meylan 1988, von Brandis et al. 2014).

On Caribbean reefs, parrotfishes are important spongivores (Dunlap & Pawlik 1996, Dunlap & Pawlik 1998). Consumption of sponges by parrotfishes has been observed following the exposure of cryptic species (Wulff 1997), through inter-habitat transplants (Dunlap & Pawlik 1996) and by juveniles (Bruggemann et al. 1994, Dunlap & Pawlik 1998, Pereira et al. 2016). In comparison to the Caribbean, evidence of sponge predation by parrotfish on Indo-Pacific reefs is scarce. In our study, 4 species of parrotfish were observed in video footage but not sampled for gut contents due to logistical difficulties encountered whilst attempting field collections. Two *Scarus ghobban* individuals were bought from the local fish market, and their gut contents were examined for evidence of sponge predation but found to contain mainly algal fragments. However, it is important to note that although parrotfish are found at the local fish market, they are not commonly targeted by local fishers but instead are caught in non-selective fish traps and fish fences, and, although we have no specific time series data regarding parrotfish biomass, a recent MSc research project found that herbivorous fish biomass increased from 2013–2017 (Sing Wong 2019). Of the 4 scarids recorded, the most numerous bites were taken by juvenile *S. flavipectoralis* (3.5% total bites), suggesting that sponge predation could be more prevalent amongst juvenile parrotfish on Wakatobi reefs. However, these juveniles appeared to be targeting mucus-bound sediment on the sponge surface as grazing scars were not evident following feeding events. This was especially apparent in the additional observations of *A. aaptos* (see the Appendix) as this sponge has a vibrant yellow interior which would be exposed following the removal of outer tissue by predation. With the exception of juvenile *S. flavipectoralis*, parrotfish represented <1% of total bites recorded on *Xestospongia* spp. Lack of representation in our data set could be explained by our choice to film open reef sponges, whereas parrotfish have been observed to rapidly consume cryptic sponges exposed by researchers (Wulff 1997). Anecdotal evidence of this type of spongivory by parrotfishes was reported from the Marshall Islands by Bakus (1967), when cryptic sponges living underneath a table *Acropora* were exposed and apparently fed upon by scarids.

Other likely spongivorous fishes were not included in the estimates of sponge consumption due to low abundances on Wakatobi reefs (e.g. pufferfishes) and time and logistical constraints (e.g. filefishes and triggerfishes). Additionally, the true diet of some prevalent sponge-grazing fishes recorded in the present study remains ambiguous due to lack of gut content

data. Four species of *Centropyge* (*bicolor*, *nox*, *tibicen*, *vroliki*) were recorded taking numerous bites from *Xestospongia* spp., and Powell et al. (2015) also recorded similar feeding behaviour in the Wakatobi MNP. Sponge was recorded as the dominant dietary item (>40% by volume) in the stomach of *C. vroliki* (Eagle & Jones 2004), and morphological studies indicate that *Centropyge* spp. are adapted to feed on attached colonial invertebrates (Konow & Bellwood 2011). However, other studies list *Centropyge* as epilithic algal grazers that are important and numerous herbivores (Green & Bellwood 2009). Similarly, *Coradion melanopus* in particular stands out as a prevalent *Xestospongia* spp. grazer, taking 18.8% of total bites but representing only 2% of sponge-grazing fish abundance. *In situ* feeding observations of *C. melanopus* at B1 suggested extremely selective grazing on barrel sponges, with 6 individuals observed to take bites almost exclusively from *Xestospongia* spp. (C. Mortimer unpubl. data). Unfortunately, gut content samples could not be collected due to small population size so sponge biomass removal by *C. melanopus* could not be verified. However, the gut contents of the congener *C. altivelis* contained 18% sponge (Nagelkerken et al. 2009), indicating that members of the *Coradion* genus may receive some nutritional benefit from associating with sponges. Future studies in the Indo-Pacific should prioritise the gut content analysis of these species to resolve their potential impact on sponge assemblages.

Acknowledgements. This research was funded by a Commonwealth doctoral scholarship from Victoria University of Wellington awarded to C.M. We thank all the volunteer and Indonesian staff at Operation Wallacea who helped facilitate this study. We are grateful to Irina Ilyushkina, Megan Shaffer and Maren Preuss who helped in the development of molecular methods and to Alberto Rovellini for field support. We are also grateful to Prof. Jonathan Gardner, Prof. Steve Wing and Dr. Chris Freeman who provided helpful comments on the work included in the manuscript. The manuscript benefitted from comments provided by Prof. Joseph Pawlik and 2 anonymous referees. Research was conducted under a number of permits from the Indonesian Ministry of Research and Technology (RISTEK) awarded to C.M. as part of a long-term collaboration with J.J. and A.H. at Hasanudin University. This project was covered by the VUW animal ethic review committee Project No. 0000024415.

LITERATURE CITED

- Abramoff MD, Magalhães PJ, Ram SJ (2004) Image processing with ImageJ. *Biophoton Int* 11:36–42
- Aronson RB, Precht WF, Toscano MA, Koltes KH (2002) The 1998 bleaching event and its aftermath on a coral reef in Belize. *Mar Biol* 141:435–447
- Bakus GJ (1967) The feeding habits of fishes and primary production at Eniwetok, Marshall Islands. *Micronesica* 3:135–149
- Bakus GJ, Nishiyama GK (2000) Three species of toxic sponges from Cebu, Philippines. *Proc Biol Soc Wash* 113: 1162–1172
- Becerro MA, Turon X, Uriz MJ, Templado J (2003) Can a sponge feeder be a herbivore? *Tylodina perversa* (Gastropoda) feeding on *Aplysina aerophoba* (Demospongiae). *Biol J Linn Soc* 78:429–438
- Bell JJ (2008) The functional roles of marine sponges. *Estuar Coast Shelf Sci* 79:341–353
- Bell JJ, Smith D (2004) Ecology of sponge assemblages (Porifera) in the Wakatobi region, south-east Sulawesi, Indonesia: richness and abundance. *J Mar Biol Assoc UK* 84:581–591
- Bell JJ, Davy SK, Jones T, Taylor MW, Webster NS (2013) Could some coral reefs become sponge reefs as our climate changes? *Glob Change Biol* 19:2613–2624
- Bell JJ, Smith D, Hannan D, Hans A, Jompa J, Thomas L (2014) Resilience to disturbance despite limited dispersal and self-recruitment in tropical barrel sponges: implications for conservation and management. *PLOS ONE* 9:e91635
- Bell JJ, Biggerstaff A, Bates T, Bennett H, Marlow J, McGrath E, Shaffer M (2017) Sponge monitoring: moving beyond diversity and abundance measures. *Ecol Indic* 78:470–488
- Bell JJ, Rovellini A, Davy SK, Taylor MW and others (2018a) Climate change alterations to ecosystem dominance: How might sponge-dominated reefs function? *Ecology* 99:1920–1931
- Bell JJ, Bennett HM, Rovellini A, Webster NS (2018b) Sponges to be winners under near-future climate scenarios. *Bioscience* 68:955–968
- Bell JJ, McGrath E, Kandler NM, Marlow J and others (2020) Inter-ocean patterns in shallow water sponge assemblage structure and function. *Biol Rev Camb Philos Soc* 95:1720–1758
- Bellwood DR, Choat JH (1990) A functional analysis of grazing in parrotfishes (family Scaridae): the ecological implications. *Environ Biol Fishes* 28:189–214
- Bellwood DR, Hughes TP, Folke C, Nyström M (2004) Confronting the coral reef crisis. *Nature* 429:827–833
- Bellwood DR, Hoey AS, Bellwood O, Goatley CHR (2014) Evolution of long-toothed fishes and the changing nature of fish–benthos interactions on coral reefs. *Nat Commun* 5:3144
- Bellwood DR, Streit RP, Brandl SJ, Tebbett SB (2019) The meaning of the term ‘function’ in ecology: a coral reef perspective. *Funct Ecol* 33:948–961
- Bennett HM, Altenrath C, Woods L, Davy SK, Webster NS, Bell JJ (2017) Interactive effects of temperature and pCO₂ on sponges: from the cradle to the grave. *Glob Change Biol* 23:2031–2046
- Berumen ML, Pratchett MS, Goodman BA (2011) Relative gut lengths of coral reef butterflyfishes (Pisces: Chaetodontidae). *Coral Reefs* 30:1005–1010
- Biggerstaff A, Smith DJ, Jompa J, Bell JJ (2015) Photo-acclimation supports environmental tolerance of a sponge to turbid low-light conditions. *Coral Reefs* 34:1049–1061
- Bruggemann JH, Vanoppen MJH, Breeman AM (1994) Foraging by the stoplight parrotfish *Sparisoma viride*. I. Food selection in different socially determined habitats. *Mar Ecol Prog Ser* 106:41–55
- Choat JH, Clements KD, Robbins WD (2002) The trophic status of herbivorous fishes on coral reefs. 1: Dietary analyses. *Mar Biol* 140:613–623

- Cleary DFR, de Voogd N (2007) Environmental associations of sponges in the Spermonde Archipelago, Indonesia. *J Mar Biol Assoc UK* 87:1669–1676
- Cleary DFR, Polónia ARM, Becking LE, de Voogd NJ, Purwanto, Gomes H, Gomes NCM (2018) Compositional analysis of bacterial communities in seawater, sediment, and sponges in the Misool coral reef system, Indonesia. *Mar Biodivers* 48:1889–1901
- Clifton J, Unsworth RKF (2013) Future directions for marine conservation in the Coral Triangle. In: Clifton J, Unsworth RKF, Smith DJ (eds) *Marine research and conservation in the Coral Triangle: the Wakatobi National Park*. Nova Science Publishers, New York, NY, p 251–258
- Cole AJ, Pratchett MS, Jones GP (2010) Corallivory in tube-lip wrasses: diet, feeding and trophic importance. *J Fish Biol* 76:818–835
- Cole AJ, Lawton RJ, Pratchett MS, Wilson SK (2011) Chronic coral consumption by butterflyfishes. *Coral Reefs* 30:85–93
- Cox EF (1986) The effects of a selective corallivore on growth rates and competition for space between two species of Hawaiian corals. *J Exp Mar Biol Ecol* 101:161–174
- Crabbe MJC, Smith DJ (2005) Sediment impacts on growth rates of *Acropora* and *Pontes* corals from fringing reefs of Sulawesi, Indonesia. *Coral Reefs* 24:437–441
- Curtis-Quick JA (2013) Drivers of change of reef fish assemblages within the Coral Triangle. PhD dissertation, The University of Essex, Colchester
- de Bakker DM, van Duyl FC, Bak RPM, Nugues MM, Nieuwland G, Meesters EH (2017) 40 Years of benthic community change on the Caribbean reefs of Curaçao and Bonaire: the rise of slimy cyanobacterial mats. *Coral Reefs* 36:355–367
- de Goeij JM, van Oevelen D, Vermeij MJA, Osinga R, Middelburg JJ, de Goeij AFPM, Admiraal W (2013) Surviving in a marine desert: the sponge loop retains resources within coral reefs. *Science* 342:108–110
- Deagle BE, Kirkwood R, Jarman SN (2009) Analysis of Australian fur seal diet by pyrosequencing prey DNA in faeces. *Mol Ecol* 18:2022–2038
- Depczynski M, Bellwood DR (2004) Microhabitat utilisation patterns in cryptobenthic coral reef fish communities. *Mar Biol* 145:455–463
- Dunlap M, Pawlik JR (1996) Video-monitored predation by Caribbean reef fishes on an array of mangrove and reef sponges. *Mar Biol* 126:117–123
- Dunlap M, Pawlik JR (1998) Spongivory by parrotfish in Florida mangrove and reef habitats. *Mar Ecol* 19:325–337
- Eagle JV, Jones GP (2004) Mimicry in coral reef fishes: ecological and behavioural responses of a mimic to its model. *J Zool (Lond)* 264:33–43
- Emslie MJ, Cheal AJ, MacNeil MA, Miller IR, Sweatman HPA (2018) Reef fish communities are spooked by SCUBA surveys and may take hours to recover. *PeerJ* 6:e4886
- Green AL, Bellwood DR (2009) Monitoring functional groups of herbivorous reef fishes as indicators of coral reef resilience: a practical guide for coral reef managers in the Asia Pacific Region. IUCN, Gland
- Gundon S, Dufayard JF, Lefort V, Anisimova M, Hordijk W, Gascuel O (2010) New algorithms and methods to estimate maximum-likelihood phylogenies: assessing the performance of PhyML 3.0. *Syst Biol* 59:307–321
- Hall TA (1999) BioEdit: a user-friendly biological sequence alignment editor and analysis program for Windows 95/98/NT. *Nucleic Acids Symp Ser* 41:95–98
- Hammond L, Wilkinson CR (1985) Exploitation of sponge exudates by coral reef holothuroids. *J Exp Mar Biol Ecol* 94:1–9
- Harmelin-Vivien ML, Bouchon-Navaro Y (1983) Feeding diets and significance of coral feeding among chaetodontid fishes in Moorea (French Polynesia). *Coral Reefs* 2:119–127
- Hay ME (1996) Marine chemical ecology: What's known and what's next? *J Exp Mar Biol Ecol* 200:103–134
- Hill MS, Lopez NA, Young KA (2005) Anti-predator defenses in western North Atlantic sponges with evidence of enhanced defense through interactions between spicules and chemicals. *Mar Ecol Prog Ser* 291:93–102
- Hoang DT, Chernomor O, von Haeseler A, Minh BQ, Vinh LS (2018) UFBBoot2: improving the ultrafast bootstrap approximation. *Mol Biol Evol* 35:518–522
- Hobson ES (1974) Feeding relationship of teleostean fishes on coral reefs in Kona, Hawaii. *Fish Bull* 72:915–1031
- Hourigan TF, Stanton FG, Motta PJ, Kelley CD, Carlson B (1989) The feeding ecology of three species of Caribbean angelfishes (Family Pomacanthidae). *Environ Biol Fishes* 24:105–116
- Kalyaanamoorthy S, Minh BQ, Wong TKF, von Haeseler A, Jermiin LS (2017) ModelFinder: fast model selection for accurate phylogenetic estimates. *Nat Methods* 14:587–589
- Konow N, Bellwood DR (2011) Evolution of high trophic diversity based on limited functional disparity in the feeding apparatus of marine angelfishes (f. Pomacanthidae). *PLOS ONE* 6:e24113
- Leray M, Agudelo N, Mills SC, Meyer CP (2013) Effectiveness of annealing blocking primers versus restriction enzymes for characterization of generalist diets: unexpected prey revealed in the gut contents of two coral reef fish species. *PLOS ONE* 8:e58076
- Leray M, Meyer CP, Mills SC (2015) Metabarcoding dietary analysis of coral dwelling predatory fish demonstrates the minor contribution of coral mutualists to their highly partitioned, generalist diet. *PeerJ* 3:e1047
- Levins R (1968) Evolution in changing environments: some theoretical explorations. *Monogr Popul Biol*, Vol 2. Princeton University Press, Princeton, NJ
- Lindel T, Hoffmann H, Hochgürtel M, Pawlik JR (2000) Structure–activity relationship of inhibition of fish feeding by sponge-derived and synthetic pyrrole-imidazole alkaloids. *J Chem Ecol* 26:1477–1496
- Marshall A, Mumby PJ (2015) The role of surgeonfish (Acanthuridae) in maintaining algal turf biomass on coral reefs. *J Exp Mar Biol Ecol* 473:152–160
- Martin D, Britayev T (1998) Symbiotic polychaetes: review of known species. *Oceanogr Mar Biol Annu Rev* 36:217–340
- McClintock JB, Amsler CD, Baker BJ, van Soest RWM (2005) Ecology of Antarctic marine sponges: an overview. *Integr Comp Biol* 45:359–368
- McGrath EC (2018) Demography and impact of habitat degradation on the giant barrel sponge *Xestospongia* spp. in the Indo-Pacific. PhD dissertation, Victoria University of Wellington
- McGrath EC, Smith DJ, Jompa J, Bell JJ (2017) Adaptive mechanisms and physiological effects of suspended and settled sediment on barrel sponges. *J Exp Mar Biol Ecol* 496:74–83
- McMurray SE, Finelli CM, Pawlik JR (2015) Population

- dynamics of giant barrel sponges on Florida coral reefs. *J Exp Mar Biol Ecol* 473:73–80
- Meylan A (1988) Spongivory in hawksbill turtles: a diet of glass. *Science* 239:393–395
- Mortimer CL (2020) Trophic interactions of marine sponges. PhD dissertation, Victoria University of Wellington
- Nagelkerken I, van der Velde G, Wartenbergh SLJ, Nugues MM, Pratchett MS (2009) Cryptic dietary components reduce dietary overlap among sympatric butterflyfishes (Chaetodontidae). *J Fish Biol* 75:1123–1143
- Nakamura H, Kobayashi J, Ohizumi Y, Hirata Y (1982) Isolation and structure of aaptamine: a novel heteroaromatic substance possessing α -blocking activity from the sea sponge *Aaptos aaptos*. *Tetrahedron Lett* 23:5555–5558
- Nielsen LA, Johnson DL (1992) Fisheries techniques. American Fisheries Society, Bethesda, MD
- Nielsen JM, Clare EL, Hayden B, Brett MT, Kratina P (2018) Diet tracing in ecology: method comparison and selection. *Methods Ecol Evol* 9:278–291
- Pawlik JR (1983) A sponge-eating worm from Bermuda: *Branchiosyllis oculata*. *Mar Ecol* 4:65–79
- Pawlik JR (2011) The chemical ecology of sponges on Caribbean reefs: natural products shape natural systems. *BioScience* 61:888–898
- Pawlik JR, Chanas B, Toonen RJ, Fenical W (1995) Defenses of Caribbean sponges against predatory reef fish. I. Chemical deterrence. *Mar Ecol Prog Ser* 127:183–194
- Pawlik JR, Loh TL, McMurray SE (2018) A review of bottom-up vs. top-down control of sponges on Caribbean fore-reefs: what's old, what's new, and future directions. *PeerJ* 6:e4343
- Pereira PHCC, Santos M, Lippi DL, Silva P (2016) Ontogenetic foraging activity and feeding selectivity of the Brazilian endemic parrotfish *Scarus zelundae*. *PeerJ* 4:e2536
- Pitt JM (1997) The feeding ecology of rabbitfish (Siganidae) at Green Island Reef: ontogenetic and interspecific differences in diet, morphology and habitat utilisation. PhD dissertation, James Cook University, Townsville
- Powell AL (2013) The impacts of predation and habitat degradation on coral reef sponge assemblages in SE Sulawesi, Indonesia. PhD dissertation, Victoria University of Wellington
- Powell A, Smith DJ, Hepburn LJ, Jones T, Berman J, Jompa J, Bell JJ (2014) Reduced diversity and high sponge abundance on a sedimented Indo-Pacific reef system: implications for future changes in environmental quality. *PLOS ONE* 9:e85253
- Powell A, Jones T, Smith DJ, Jompa J, Bell JJ (2015) Spongivory in the Wakatobi Marine National Park, southeast Sulawesi, Indonesia. *Pac Sci* 69:487–508
- Pratchett MS (2005) Dietary overlap among coral-feeding butterflyfishes (Chaetodontidae) at Lizard Island, northern Great Barrier Reef. *Mar Biol* 148:373–382
- Purcell SW, Bellwood DR (1993) A functional analysis of food procurement in two surgeonfish species, *Acanthurus nigroluscus* and *Ctenochaetus striatus* (Acanthuridae). *Environ Biol Fishes* 37:139–159
- Randall JE, Hartman WD (1968) Sponge-feeding fishes of the West Indies. *Mar Biol* 1:216–225
- Rix L, de Goeij JM, van Oevelen D, Struck U, Al-Horani FA, Wild C, Naumann MS (2018) Reef sponges facilitate the transfer of coral-derived organic matter to their associated fauna via the sponge loop. *Mar Ecol Prog Ser* 589:85–96
- Rovellini A, Dunn MR, Fulton EA, Webster NS and others (2019) Decadal variability in sponge abundance and biodiversity on an Indo-Pacific coral reef. *Mar Ecol Prog Ser* 620:63–76
- RStudio Team (2015) RStudio: integrated development for R. RStudio, PBC, Boston, MA
- Ruzicka RR, Colella MA, Porter JW, Morrison JM and others (2013) Temporal changes in benthic assemblages on Florida Keys reefs 11 years after the 1997/1998 El Niño. *Mar Ecol Prog Ser* 489:125–141
- Sano M (1989) Feeding habits of Japanese butterflyfishes (Chaetodontidae). *Environ Biol Fishes* 25:195–203
- Schrader C, Schielke A, Ellerbroek L, John R (2012) PCR inhibitors — occurrence, properties and removal. *J Appl Microbiol* 113:1014–1026
- Sing Wong A (2019) Functional structure of herbivorous fish communities in the Wakatobi National Park, Southeast Sulawesi. MSc dissertation, The University of Essex, Colchester
- Stefanoudis PV, Gress E, Pitt JM, Smith SR and others (2019) Depth-dependent structuring of reef fish assemblages from the shallows to the rariphotic zone. *Front Mar Sci* 6:307
- Symondson WOC (2002) Molecular identification of prey in predator diets. *Mol Ecol* 11:627–641
- van Soest RWM, Hooper JNA (2002) Order Haplosclerida Topsent, 1928. In: Hooper JNA, van Soest RWM, Willenz P (eds) *Systema Porifera: a guide to the classification of sponges*. Springer, Boston, MA, p 831–832
- van Soest RWM, Boury-Esnault N, Vacelet J, Dohrmann M and others (2012) Global diversity of sponges (Porifera). *PLOS ONE* 7:e35105
- Venables WN, Ripley BD (2002) *Modern applied statistics with S*, 4th edn. Springer, New York, NY
- Verdin Padilla CJ, Carballo JL, Camacho ML (2010) A qualitative assessment of sponge-feeding organisms from the Mexican Pacific coast. *Open Mar Biol J* 4:39–46
- Vestheim H, Jarman SN (2008) Blocking primers to enhance PCR amplification of rare sequences in mixed samples — a case study on prey DNA in Antarctic krill stomachs. *Front Zool* 5:12
- Vicente J, Silbiger NJ, Beckley BA, Raczkowski CW, Hill RT (2016) Impact of high $p\text{CO}_2$ and warmer temperatures on the process of silica biomineralization in the sponge *Mycale grandis*. *ICES J Mar Sci* 73:704–714
- von Brandis RG, Mortimer JA, Reilly BK, van Soest RWM, Branch GM (2014) Diet composition of hawksbill turtles (*Eretmochelys imbricata*) in the Republic of Seychelles. *West Indian Ocean J Mar Sci* 13:81–91
- Wild C, Huettel M, Klueber A, Kremb SG, Rasheed MYM, Jørgensen BB (2004) Coral mucus functions as an energy carrier and particle trap in the reef ecosystem. *Nature* 428:66–70
- Wilson DM, Puyana M, Fenical W, Pawlik JR (1999) Chemical defense of the Caribbean reef sponge *Axinella corrugata* against predatory fishes. *J Chem Ecol* 25:2811–2823
- Wulff JL (1997) Parrotfish predation on cryptic sponges of Caribbean coral reefs. *Mar Biol* 129:41–52
- Wulff JL (2006) Ecological interactions of marine sponges. *Can J Zool* 84:146–166
- Wulff J (2017) Bottom-up and top-down controls on coral reef sponges: disentangling within-habitat and between-habitat processes. *Ecology* 98:1130–1139
- Zoltowska-Aksamitowska S, Tsurkan MV, Lim SC, Meissner H and others (2018) The demosponge *Pseudoceratina purpurea* as a new source of fibrous chitin. *Int J Biol Macromol* 112:1021–1028

Appendix. Additional video observations

Petrosia corticata at The Ridge

During the 148 min of footage recorded on the exposed inner tissue of *Petrosia corticata*, 888 bites were recorded giving a grazing intensity rate of $112.6 \text{ min}^{-1} \text{ m}^{-2}$. The majority of bites were directed at an area of exposed choanosome, and sponge tissue was observed being removed by all species. The most voracious predator was *Zanclus cornutus* which took 59.5% of bites, followed by *Pygoplites diacanthus* taking 25.5% and *Siganus punctatus* at 6.6%. Other species observed taking bites included *Pomacanthus imperator*, *Heniochus singularis*, *Centropyge nox* and *Scarus ghobban* (terminal phase). *P. imperator* took the highest no. of bites in a single feeding event (34).

Lamellodysidea herbacea at Sampela

In the 130 minutes of footage recorded on *L. herbacea*, 68 bites were observed on the sponge surface giving a total grazing intensity rate of $4.4 \text{ min}^{-1} \text{ m}^{-2}$. *Scarus flavipectoralis* juveniles took the most numerous bites (45.6%), followed by *Ctenochaetus binotatus* (29.4%). *Z. cornutus* and *Centropyge kleinii* took 8 and 7 bites respectively and 2 bites were made by *Centropyge tibicen*. Although bites were recorded on sponge surface, it was not immediately evident that sponge tissue was being removed.

Aaptos suberitoides at Sampela

In the 230 min of footage recorded on *Aaptos suberitoides*, 133 bites were recorded giving a grazing intensity rate of $4.6 \text{ min}^{-1} \text{ m}^{-2}$. The most numerous bites recorded by the angelfish *Centropyge bicolor* (76.7%) followed by the juvenile parrotfish *S. flavipectoralis* (9.8%); $\leq 5\%$ of bites were recorded by *Scarus niger*, *C. tibicen* and *Z. cornutus*. *Canthigaster valentini*, *C. kleinii*, *Halichoeres prosopeton* and *Sufflamen bursa* all took only one bite in the whole video.

Tubular sponge (possibly *Theonella* sp.) at KDS

In the 33 min of footage recorded on the green tubular sponge, a total of 385 bites were recorded by 6 species. The majority of bites were taken by *P. diacanthus* (295) and were directed towards an exposed bit of choanosome which made up approximately 4.8% of total sponge area. This area of exposed inner tissue had a grazing intensity rate from *P. diacanthus* of $265.8 \text{ bites min}^{-1} \text{ m}^{-2}$ whereas the undisturbed sponge surface had a grazing intensity rate of $10.3 \text{ bites min}^{-1} \text{ m}^{-2}$. *Z. cornutus* took the second highest number of bites (47) all of which were directed towards the white exposed inner tissue giving a grazing intensity rate of $74.8 \text{ bites min}^{-1} \text{ m}^{-2}$. *Pomacanthus xanthurus* was also sighted feeding from the sponge, taking 4 bites on the white exposed inner tissue. Other species noted were *C. tibicen*, *C. kleinii* and *C. binotatus*, taking 25, 9 and 4 bites respectively, all from the undisturbed sponge surface.

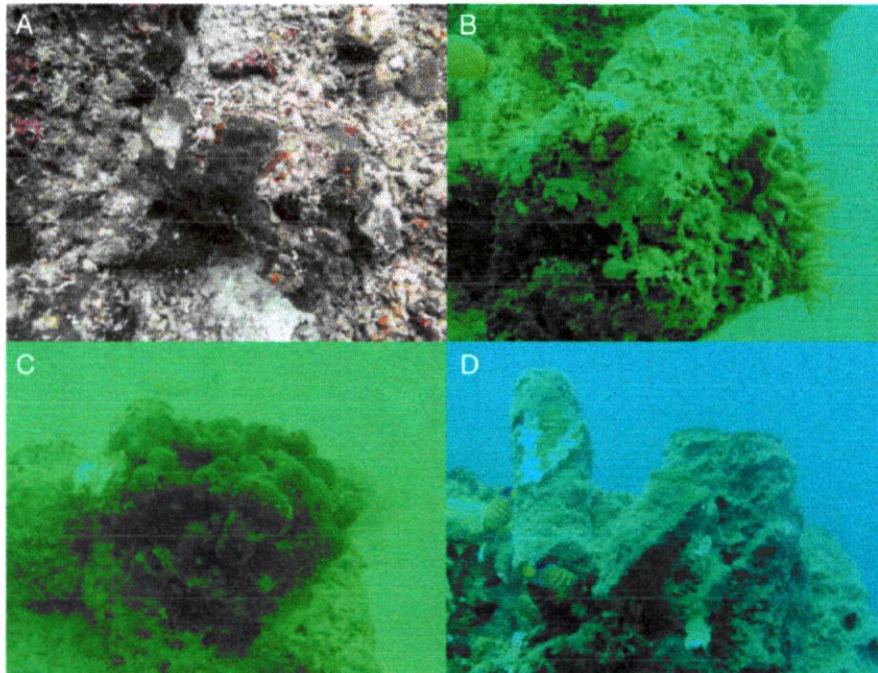


Fig. A1. Screenshots of (A) *Petrosia corticata*, (B) *Lamellodysidea herbacea*, (C) *Aaptos suberitoides* (D) green tubular sponge

Editorial responsibility: Tim McClanahan
Mombasa, Kenya
Reviewed by: J. Luis Carballo and 2 anonymous referees

Submitted: November 20, 2020
Accepted: June 8, 2021
Proofs received from author(s): August 14, 2021

MEPS_2021_Charlotte_Montimer_Jamaluddin_Jompa_Abdul_...

ORIGINALITY REPORT

11 %
SIMILARITY INDEX

8 %
INTERNET SOURCES

7 %
PUBLICATIONS

2 %
STUDENT PAPERS

PRIMARY SOURCES

1 researcharchive.vuw.ac.nz 1 %
Internet Source

2 link.springer.com 1 %
Internet Source

3 Peter J. Edmunds, Matthew Coblenz, Janie Wulff. "A quarter-century of variation in sponge abundance and community structure on shallow reefs in St. John, US Virgin Islands", *Marine Biology*, 2020 1 %
Publication

4 epubs.aims.gov.au <1 %
Internet Source

5 Emily C. McGrath, David J. Smith, Jamaluddin Jompa, James J. Bell. "Adaptive mechanisms and physiological effects of suspended and settled sediment on barrel sponges", *Journal of Experimental Marine Biology and Ecology*, 2017 <1 %
Publication

6 www.bioone.org <1 %
Internet Source

| | | |
|----|---|------|
| 7 | espace.library.uq.edu.au Internet Source | <1 % |
| 8 | Submitted to Turun yliopisto Student Paper | <1 % |
| 9 | peerj.com Internet Source | <1 % |
| 10 | www.nature.com Internet Source | <1 % |
| 11 | orca.cf.ac.uk Internet Source | <1 % |
| 12 | onlinelibrary.wiley.com Internet Source | <1 % |
| 13 | Submitted to University of St Andrews Student Paper | <1 % |
| 14 | dez.pensoft.net Internet Source | <1 % |
| 15 | SE McMurray, AD Stubler, PM Erwin, CM Finelli, JR Pawlik. "A test of the sponge-loop hypothesis for emergent Caribbean reef sponges", Marine Ecology Progress Series, 2018 Publication | <1 % |
| 16 | www.frontiersin.org Internet Source | <1 % |

might sponge-dominated reefs function?",
Ecology, 2018

Publication

-
- **25** Submitted to UC, Irvine <1 %
Student Paper

 - **26** "Climate Change, Ocean Acidification and Sponges", Springer Science and Business Media LLC, 2017 <1 %
Publication

 - 27** scholarspace.manoa.hawaii.edu <1 %
Internet Source

 - 28** www.tandfonline.com <1 %
Internet Source

 - **29** complete.bioone.org <1 %
Internet Source

 - 30** people.uncw.edu <1 %
Internet Source

 - 31** s3.amazonaws.com <1 %
Internet Source

 - 32** www.opwall.com <1 %
Internet Source

 - **33** James J. Bell, Emily McGrath, Nora M. Kandler, Joseph Marlow et al. "Interocean patterns in shallow water sponge assemblage structure and function", Biological Reviews, 2020 <1 %
Publication
-

| | | |
|----|--|------|
| 34 | benthamopen.com Internet Source | <1 % |
| 35 | d-nb.info Internet Source | <1 % |
| 36 | www.chemie.uni-muenchen.de Internet Source | <1 % |
| 37 | www.irjponline.com Internet Source | <1 % |
| 38 | www.scielo.br Internet Source | <1 % |
| 39 | A. Biggerstaff, D. J. Smith, J. Jompa, J. J. Bell. "Photoacclimation supports environmental tolerance of a sponge to turbid low-light conditions", Coral Reefs, 2015 Publication | <1 % |
| 40 | Alberto Rovellini, Matthew R. Dunn, Elizabeth A. Fulton, Lisa Woods, Jamaluddin Jompa, Abdul Haris, James J. Bell. "Interannual variability and decadal stability of benthic organisms on an Indonesian coral reef", Journal of the Marine Biological Association of the United Kingdom, 2021 Publication | <1 % |
| 41 | Bell, James J., Emily McGrath, Andrew Biggerstaff, Tracey Bates, Holly Bennett, Joseph Marlow, and Megan Shaffer. | <1 % |

"Sediment impacts on marine sponges",
Marine Pollution Bulletin, 2015.

Publication

-
- **42** Submitted to Imperial College of Science, Technology and Medicine <1 %
Student Paper

 - **43** journals.plos.org <1 %
Internet Source

 - **44** nbn-resolving.org <1 %
Internet Source

 - **45** I. Nagelkerken. "Cryptic dietary components reduce dietary overlap among sympatric butterflyfishes (Chaetodontidae)", Journal of Fish Biology, 10/2009 <1 %
Publication

 - **46** Inka Vanwonderghem, Nicole S. Webster. "Coral Reef Microorganisms in a Changing Climate", iScience, 2020 <1 %
Publication

 - **47** Janie L. Wulff. "Targeted predator defenses of sponges shape community organization and tropical marine ecosystem function", Ecological Monographs, 2020 <1 %
Publication

 - **48** MS Hill, NA Lopez, KA Young. "Anti-predator defenses in western North Atlantic sponges with evidence of enhanced defense through

interactions between spicules and chemicals",
Marine Ecology Progress Series, 2005

Publication

-
- **49** Robert H. Reavis, Joshua M. Copus. <1 %
"Monogamy in a feeding generalist,
Chaetodon trichrous, the endemic Tahitian
Butterflyfish", Environmental Biology of
Fishes, 2011
Publication

 - **50** Submitted to University of Sydney <1 %
Student Paper

 - **51** archive.org <1 %
Internet Source

 - **52** elib.suub.uni-bremen.de <1 %
Internet Source

 - **53** hdl.handle.net <1 %
Internet Source

 - **54** nsuworks.nova.edu <1 %
Internet Source

 - **55** researchonline.jcu.edu.au <1 %
Internet Source

 - **56** uknowledge.uky.edu <1 %
Internet Source

 - **57** www.observadorsdelmar.cat <1 %
Internet Source

58 A. Biggerstaff, J. Jompa, J. J. Bell. "Increasing benthic dominance of the phototrophic sponge *Lamellodysidea herbacea* on a sedimented reef within the Coral Triangle", *Marine Biology*, 2017
Publication

59 Ruzicka, R.. "Sponge community structure and anti-predator defenses on temperate reefs of the South Atlantic Bight", *Journal of Experimental Marine Biology and Ecology*, 20091115
Publication

60 A Rovellini, MR Dunn, EA Fulton, NS Webster, DJ Smith, J Jompa, A Haris, J Berman, JJ Bell. "Decadal variability in sponge abundance and biodiversity on an Indo-Pacific coral reef", *Marine Ecology Progress Series*, 2019
Publication

61 James J. Bell, Andrew Biggerstaff, Tracey Bates, Holly Bennett, Joseph Marlow, Emily McGrath, Megan Shaffer. "Sponge monitoring: Moving beyond diversity and abundance measures", *Ecological Indicators*, 2017
Publication

62 Megan R. Shaffer, Heidi M. Luter, Nicole S. Webster, Muhammad A. Abdul Wahab, James J. Bell. "Evidence for genetic structuring and

limited dispersal ability in the Great Barrier Reef sponge *Carteriospongia foliascens*", Coral Reefs, 2019

Publication

63

eprints.utas.edu.au

Internet Source

<1 %

Exclude quotes On

Exclude matches < 5 words

Exclude bibliography On

| | | |
|----|---|------|
| 17 | A. J. Cole, R. J. Lawton, M. S. Pratchett, S. K. Wilson. "Chronic coral consumption by butterflyfishes", Coral Reefs, 2010 Publication | <1 % |
| 18 | nbn-resolving.de Internet Source | <1 % |
| 19 | nesptropical.edu.au Internet Source | <1 % |
| 20 | epubs.scu.edu.au Internet Source | <1 % |
| 21 | escholarship.org Internet Source | <1 % |
| 22 | DV Fairclough. "Partitioning of marine transition zone reefs among temperate, subtropical and tropical fishes is related more to depth and habitat than temperature", Marine Ecology Progress Series, 2021 Publication | <1 % |
| 23 | James J Bell, Holly M Bennett, Alberto Rovellini, Nicole S Webster. "Sponges to Be Winners under Near-Future Climate Scenarios", BioScience, 2018 Publication | <1 % |
| 24 | James J. Bell, Alberto Rovellini, Simon K. Davy, Michael W. Taylor et al. "Climate change alterations to ecosystem dominance: how | <1 % |